Directive 98/8/EC concerning the placing of biocidal products on the market

Inclusion of active substances in Annex I or IA to Directive 98/8/EC

Assessment Report



Copper (II) oxide

Product-type 8 (Wood preservatives)

September 2011

Annex I - France

Copper (II) oxide (PT 8)

Assessment report

FINALISED IN THE STANDING COMMITTEE ON BIOCIDAL PRODUCTS AT ITS MEETING ON 22/09/2011 IN VIEW OF ITS INCLUSION IN ANNEX I TO DIRECTIVE 98/8/EC

CONTENTS

| 1 - STATEMENT | OF SUBJECT MATTER AND PURPOSE | 3 |
|--------------------|---|---|
| PURPOSE OF THE | ASSESSMENT REPORT | 3 |
| OVERALL CONCL | USION IN THE CONTEXT OF DIRECTIVE 98/8/EC | |
| | MMARY AND CONCLUSIONS | |
| 2.1 GENERA | AL SUBSTANCE INFORMATION / GENERAL PRODUCT INFORMATION | 3 |
| | tity and physico-chemical properties of the active substance | |
| | tity and physico-chemical properties of the biocidal product | |
| | acy and intended usesacy and intended uses | |
| | sification and labelling | |
| | RY OF THE RISK ASSESSMENT | |
| | | |
| | an health risk assessment | 3 |
| 2.2.1.1 2.2.1.2 | Hazard identification and effects assessment | |
| 2.2.1.2 | Exposure assessment | |
| | ronment risk assessment | |
| | | |
| 2.2.2.1 2.2.2.2 | Fate and distribution in the environment | 3 |
| 2.2.2.2 | Environmental effect assessment | |
| 2.2.2.2.1 | Freshwater compartment Sediment compartment | |
| 2.2.2.2.3 | Terrestrial compartment | |
| 2.2.2.2.4 | STP compartment | |
| 2.2.2.2.5 | Summary of PNECs | |
| 2.2.2.3 | Risk characterisation for the environment | |
| 2.2.2.3.1 | Aquatic compartment (including sediment) | |
| 2.2.2.3.2 | Sewage treatment plant organism | |
| 2.2.2.3.3 | Atmosphere | 3 |
| 2.2.2.3.4 | Terrestrial compartment | |
| 2.2.2.3.5 | Groundwater | |
| 2.2.2.3.6 | Non compartment specific effects relevant to the food chain (secondary poisoning) | |
| 2.2.2.4 | PBT assessment | |
| 3 – PROPOSAL F | OR THE DECISION | |
| | D TO THE PROPOSED DECISION | |
| | ECISION REGARDING THE INCLUSION IN ANNEX I OR IA | |
| | E TAKEN INTO ACCOUNT BY MEMBER STATES WHEN AUTHORISING PRODUCTS | |
| | | |
| | FURTHER INFORMATION | |
| APPENDIX 1 – LI | STING OF ENDPOINTS | 3 |
| APPENDIX II: LI | ST OF INTENDED USES | 3 |
| ADDENDIVITE | ICT OF CTUDIES | 2 |
| APPENDIX III: L | JIST OF STUDIES | 3 |

1 - STATEMENT OF SUBJECT MATTER AND PURPOSE

This assessment report has been established as a result of the evaluation of copper (II) oxide as product-type 8 (wood preservative), carried out in the context of the work programme for the review of existing active substances provided for in Article 16(2) of Directive 98/8/EC concerning the placing of biocidal products on the market, with a view to the possible inclusion of this substance into Annex I to the Directive.

Copper oxide (CAS no 1317-38-0) was notified as an existing active substance, by the Wood Preservative Task Force, hereafter referred to as the applicant, in product-type 8.

Commission Regulation (EC) No 1451/2007 of 4 December 2007 lays down the detailed rules for the evaluation of dossiers and for the decision-making process in order to include or not an existing active substance into Annex I or IA to the Directive.

In accordance with the provisions of Article 7(1) of that Regulation, the Commission designated France as Rapporteur Member State to carry out the assessment of copper oxide on the basis of the dossier submitted by the applicant. The deadline for submission of a complete dossier for copper oxide as an active substance in product-type was 28 March 2004, in accordance with Article 9(2) of Regulation (EC) No 1451/2007.

On 29 March 2004, the French competent authority received a dossier from the applicant. The Rapporteur Member State accepted the dossier as complete for the purpose of the evaluation, taking into account the supported uses, and confirmed the acceptance of the dossier on 28 September 2004.

On 10 May 2007, the Rapporteur Member State submitted, in accordance with the provisions of Article 14(4) and (6) of Regulation (EC) No 1451/2007, to the Commission and the applicant a copy of the evaluation report, hereafter referred to as the competent authority report. The Commission made the report available to all Member States by electronic means on 23 July 2007. The competent authority report included a recommendation for the inclusion of copper (II) oxide in Annex I to the Directive for product-type 8.

In accordance with Article 16 of Regulation (EC) No 1451/2007, the Commission made the competent authority report publicly available by electronic means on 9 April 2008. This report did not include such information that was to be treated as confidential in accordance with Article 19 of Directive 98/8/EC.

In order to review the competent authority report and the comments received on it, consultations of technical experts from all Member States (peer review) were organised by the Commission. Revisions agreed upon were presented at technical and competent authority meetings and the competent authority report was amended accordingly.

On the basis of the final competent authority report, the Commission proposed the inclusion of copper (II) oxide in the Annex I of Directive 98/8/EC and consulted the Standing Committee on Biocidal Products on 22/09/2011.

In accordance with Article 15(4) of Regulation (EC) No 1451/2007, the present assessment report contains the conclusions of the Standing Committee on Biocidal Products, as finalised during its meeting held on 22/09/2011.

¹ Directive 98/8/EC of the European Parliament and of the Council of 16 February 1998 concerning the placing biocidal products on the market, OJ L 123, 24.4.98, p.1

_

² Commission Regulation (EC) No 1451/2007 of 4 December 2007 on the second phase of the 10-year work programme referred to in Article 16(2) of Directive 98/8/EC of the European Parliament and of the Council concerning the placing of biocidal products on the market. OJ L 325, 11.12.2007, p. 3

Purpose of the assessment report

This assessment report has been developed and finalised in support of the decision to include copper (II) oxide in the Annex I of Directive 98/8/EC for product-type 8.

The applicant is not currently placing nano forms of copper oxide on the market. Therefore, the submitted dossier and the finalised assessment report don't cover potential nanoforms of this copper compound, should such forms exist.

The aim of the assessment report is to facilitate the authorisation in Member States of individual biocidal products in product-type 8 that contain copper (II) oxide. In their evaluation, Member States shall apply the provisions of Directive 98/8/EC, in particular the provisions of Article 5 as well as the common principles laid down in Annex VI.

For the implementation of the common principles of Annex VI, the content and conclusions of this assessment report, which is available at the Commission website³, shall be taken into account.

However, where conclusions of this assessment report are based on data protected under the provisions of Directive 98/8/EC, such conclusions may not be used to the benefit of another applicant, unless access to these data has been granted.

Overall conclusion in the context of Directive 98/8/EC

The overall conclusion from the evaluation is that it may be expected that there are products containing copper (II) oxide for the product-type 8, which will fulfil the requirements laid down in Article 5 of Directive 98/8/EC. This conclusion is however subject to:

- i. compliance with the particular requirements in the following sections of this assessment report,
- ii. the implementation of the provisions of Article 5(1) of Directive 98/8/EC, and
- iii. the common principles laid down in Annex VI to Directive 98/8/EC.

Furthermore, these conclusions were reached within the framework of the uses that were proposed and supported by the applicant (see Appendix II). Extension of the use pattern beyond those described will require an evaluation at product authorisation level in order to establish whether the proposed extensions of use will satisfy the requirements of Article 5(1) and of the common principles laid down in Annex VI to Directive 98/8/EC.

³ <u>http://ec.europa.eu/comm/environment/biocides/index.htm</u>

2 – OVERALL SUMMARY AND CONCLUSIONS

2.1 General substance information / general product information

Copper (II) oxide has been evaluated for its use as wood preservative (product type 8) by professionals. Considering that cupric ion is the real active substance, the risk assessment was not only based on copper oxide data but also on other salts data (mainly on those of copper sulphate).

In other respects, as all the wood preservative copper-based products placed on the EU market also contain other active substance, it was chosen to present an amine copper formulation as the most representative biocidal product. In the case of copper/amine product type, other active substances are included in the final formulation. These other active substances can include quaternary ammonium compounds, tebuconazole, propiconazole, boric acid... This list is not exhaustive and other active substances may be included.

For the purposes of Annex I listing, it is considered that only the copper content of the product should be reviewed at this stage and at the time of product authorisation, a full dossier will be provided by the formulator parties of the Wood Preservative Copper Task Force.

The applicant is not currently placing nano forms of copper oxide on the market. Therefore, the submitted dossier and the finalised assessment report don't cover potential nanoforms of this copper compound, should such forms exist.

The main identity and the physical/chemical properties of Copper Oxide are given in Appendix 1 (Listing of endpoints). The active substance shall comply with the specification given in Appendix 1 of this report. The evaluation has established that for the active substance notified by the Wood Preservative Task Force, traces of metals of toxicological concern (arsenic, lead, cadmium and nickel) were identified among the manufacturing impurities

2.1.1 Identity and physico-chemical properties of the active substance

The active substance, such as defined by the Biocidal Directive, is copper (II) oxide (CuO) (CAS no. 1317-38-0), with a minimum purity of 97.6%. It is also known as cupric oxide.

In contrast, it is acknowledged that the actual active substance (i.e. the chemical part really involved in the physiological processes) is the cupric ion (Cu²⁺), released by copper oxide when in contact with biological media. Traces of metals of toxicological concern (arsenic, cadmium, lead and nickel) were identified among the manufacturing impurities.

Three sources of copper (II) oxide out of four submitted in the dossier are accepted. See the confidential part for the specifications. However, to confirm the data presented in the dossier, a new 5-batch analysis is required at the product authorization stage to check the compliance of current production of each source to these specifications.

Copper oxide is a dark grey powder with no characteristic odour and a molecular weight of 79.55 g/mol. It has a relative density of 1.018 and a melting point of 1446°C. Its vapour pressure is not measurable due to its high melting point. It is slightly soluble in water (solubility in water at pH 6 (at 20°C) = 3.94×10^{-4} g/L). It is neither flammable nor explosive nor oxidizing.

Adequate analytical methods are available for the determination of copper and the known impurities in the active substance as manufactured. These methods were collaborately validated and are very widely used, but validation data on the active substance must be provided. It must be highlighted that methods

of analysis for one impurity > 0.1% was not provided and must be provided at the product authorization stage.

The analyses of copper in environmental matrices and body fluids and tissues are routinely performed in many laboratories. As these methods were collaborately validated and are very widely used, limited validation data were accepted.

2.1.2 Identity and physico-chemical properties of the biocidal product

| Trade name | Amine Copper Solution/Product | | | |
|-------------------------------|-------------------------------|--------------------|--|--|
| Manufacturer's development | Not applicable | | | |
| code number(s) | | | | |
| Ingredient of preparation | Function | Content | | |
| Copper (from copper | Active substance | Typical conc.: 13% | | |
| carbonate or copper oxide) | | | | |
| 2-aminoethanol | Solubilizing agent | Up to 40% | | |
| Water | Co-formulant | Make up to 100% | | |
| Physical state of preparation | Liquid | | | |
| Nature of preparation | Water based formulation | | | |

Concerning the biocidal product, the amine copper solution is an intense cobalt blue liquid, with a slight ammonia odour. Its pH is slightly basic since pH = 9.54 at 23°C. It has a relative density of 1.220. It has neither flammable nor explosive properties. It is a weak oxidising agent.

2.1.3 Intended uses and efficacy

• Field of use / Function / Mode of action

Field of use

Product type 8 (PT): wood preservative

Copper oxide is intended to be used as a preventive wood preservative for wood in Use class 1, 2, 3 4-1 and 4-2, as defined in the EN 335⁴.

The active substance is restricted to industrial use only, in timber treatment plants operated by trained personnel. The product is supplied as a water-soluble concentrate that is mixed with clean water in the industrial timber treatment plant concentrate storage, product mixing and dilution system. The product is actually applied in simple vacuum pressure timber impregnation plants for the treatment of timbers. Formulated wood preservatives contain 0.317% copper for the Use Class 3 and 0.57% copper for the Use Class 4 in the ready-to-use product, knowing that the biocidal product containing the active substance presented in this dossier (ACQ-C2D) is an amine copper product.

o Function

The active substance Copper (II) oxide acts as a fungicide and as an insecticide use for preventive wood preservation (product type 8).

⁴ Since 2007 and the revision of the EN335-1, use classes had replaced hazard classes.

o Mode of action

As the active substance is the Cu²⁺ ion, copper oxide is therefore described as the precursor to release of the cupric ion. As a consequence, most copper-containing formulations are described in terms of total copper.

Fungi:

It is considered that the fungicidal properties of copper compounds are dependent on the affinity of the copper ion (Cu²⁺) for different chemical groups within cells, particularly thiol groups, resulting in the non-specifc denaturation of proteins and enzymes. In addition, it is thought that the ion can interfere with the activity of the pyruvate dehydrogenase system inhibiting the conversion of pyruvate to acetyl CoA within mitochondria. Copper reacts with most essential elements within a cell. It also reacts with ligands on the cell surface and this can interfere with membrane function. Copper may also act extracellulary in the case of fungi and inhibit the production of fungal extracellular enzymes.

Insects:

Copper in toxic doses acts as a stomach poison.

Termites: copper ions act on the gut symbionts killing the gut microflora and fauna and depriving the termite of its ability to digest cellulose.

• Objects to be protected, target organism

The fact is that no single active substance based wood preservative formulations are placed on the market because of the range of organisms to be controlled. As the exact composition of a formulation depends on the end use of the treated timber, a reading across of data had to be done in assessing the efficacy of the active substance.

Copper efficacy was examined for the following target organisms:

| Application mode | Target organism | Active substances rate |
|-------------------------------------|--|---|
| Vacuum pressure timber impregnation | Fungi: Wood rotting basidiomycetes and soft rot fungi | The typical use of copper for each use class is likely to be: UC 1,2,3: ≤ 1.9 kg Cu /m ³ |
| | Insects: Wood boring beetles and termites. | sapwood loading UC 4: ≤ 3.42 kg Cu / m³ sapwood |
| | There is no claim concerning the efficacy of copper on blue stain and mould, although efficacy against these organisms is well known and documented. | loading. These data are based on the current state of the art concerning the practical uses of copper-based products during the last decades, related to the expected |
| | documentod. | service life of wooden elements treated with copper-based product. |

Studies provided with tests on Cu²⁺ delivered as copper sulphate or as copper carbonate. The provided data are partial but sufficient to efficacy of copper against target organisms, even it trials provided partially meet EN 559 requirements.

As copper based wood preservative is used in conjunction with other biocides, full efficacy data should be provided by the individual applicants at product authorisation stage.

• Resistance

According to the data submitted, there are strains of some species of wood destroying fungi that exhibit tolerance to copper. Additional biocides are used where necessary to control copper-tolerant strains of fungi.

According to the data submitted, there is no evidence of insects being naturally tolerant or being able to develop resistance to copper at the level of copper used for biocidal purposes in wood preservation.

2.1.4 Classification and labelling

On the basis of a review of submitted data, the following classification and labelling is proposed:

• Active substance

| | Directive 67/548/EEC | | | | |
|---|---|---|--|--|--|
| Class of danger | Xn - Harmful | | | | |
| | N - Dange | rous for the environment | | | |
| R phrases | R20 Harm | ful by inhalation | | | |
| | | Very toxic to aquatic organisms, may cause long-term adverse effects in a environment | | | |
| S phrases | S22: Do not breathe in dust | | | | |
| | S60: This | material and its container must be disposed of as hazardous waste | | | |
| | S61: Avoid release to the environment. Refer to special instructions/safety data sheets | | | | |
| | | Regulation 1272/2008 | | | |
| Hazard classes and car | tegories / | Acute Tox. 4 /H332 – Harmful if inhaled | | | |
| hazard statements | | Aquatic Acute/H400 – Very toxic to aquatic life | | | |
| Aquatic chronic/H410 - Very toxic to aquatic life with long lasting effects | | | | | |

• Biocidal product

| Directive 67/548/EEC | | | | | |
|----------------------|--|--|--|--|--|
| Class of danger | C – Corrosive T - Toxic | | | | |
| R phrases | R22:Harmful if swallowed R23: Toxic by inhalation R34: Causes burns R51: Toxic to aquatic organisms R53: May cause long-term adverse effects in the aquatic environment. | | | | |
| S phrases | S25:Avoid contact with eyes S28:After contact with skin, wash immediately with plenty of water S36/37/39:Wear suitable protective clothing, gloves and eye/face protection S38:In case of insufficient ventilation wear suitable respiratory equipment S45:In case of an accident or if you fell unwell seek medical S60: This material and its container must be disposed of as hazardous waste. S61: Avoid release to the environment. Refer to special instructions/safety data sheets. | | | | |
| | Regulation 1272/2008 | | | | |

| Copper (II) oxide | 22/09/2011 |
|-------------------|------------|
| | |

| | Acute Tox. 4 /H302: Harmful if swallowed Acute Tox. 3 /H331: Toxic if inhaled Skin Corr. 1B/H314: Causes skin burns and eye damages Aquatic Chronic /H411: Toxic to aquatic life with long lasting effects. |
|--|---|
|--|---|

2.2 Summary of the risk assessment

2.2.1 Human health risk assessment

2.2.1.1 HAZARD IDENTIFICATION AND EFFECTS ASSESSMENT

Foreword

Copper is an essential metal for life and is employed in all human cells. The main daily dietary intake of copper in adults ranges between 1.5 and 3.0 mg (IPCS, 1998). Most human diets naturally contain between 1 and 2 mg/person/day of copper, with some containing up to 4 mg/person/day.

Copper is regulated by homeostatic mechanism. Homeostasis can be described as the maintenance of a constant internal environment in response to changes in internal and external environments. Homeostatic maintenance requires the tightly coordinated control of copper uptake, distribution and efflux in cells and the organism as a whole. The ability of the body to control the uptake and excretion of copper makes this an important factor in considering the exposure and effects of essential elements like copper.

Copper is involved in the reactions and functions of many enzymes, including angiogenesis, neurohormone release, oxygen transport and regulation of genetic expression. In this scope, copper carbonate can be considered as a precursor, releasing cupric ion, which is the actual active substance. This explains that, while several endpoints were documented by studies directly performed with copper carbonate (acute toxicity, skin and eye irritation, sensitisation), other endpoints were documented by other copper salts (mainly copper sulphate).

In mammalian toxicity, it is also considered that the most toxic moiety of any copper salt is the Cu²⁺ ion. This can be shown through the comparison of the most soluble salt (copper sulphate) with other relatively insoluble copper salts, where the solubility, bioavailability and hence toxicity of these salts can vary – with copper sulphate representing the worst-case scenario. When the acute oral toxicity of this salt is compared with either copper oxide or copper carbonate, the data indicate that copper sulphate is more toxic and thus more bioavailable. Therefore all the properties described below for copper will also be applicable to copper (II) oxide.

This has also been confirmed in comparative bioavailability studies where copper sulphate was shown to be more or equally bioavailable in relation to the copper carbonate in poultry and swine.

Moreover, as presented in the table below (Table 2.2-1) the Copper sulphate is more toxic than the other copper compounds, and then using studies performed with copper sulphate could be considered as a worst-case.

Table 2.2.1.1-1: Comparative toxicity of the different Copper Salts

| Copper | Colubility | Acute toxic | Irritation | |
|---------------------|-------------------------|-------------|-------------|------------|
| salts | Solubility | oral | dermal | irritation |
| CuSO ₄ | 317 g/L 482 mg/kg >1000 | | >1000 mg/kg | R36/38 |
| CuCO ₃ | 1,5 mg/L | 1400 mg/kg | >2000 mg/kg | NC |
| CuO | 0,3 mg/L | >2000 mg/kg | >2000 mg/kg | NC |
| Cu(OH) ₂ | 6.6μ/L | 763 mg/kg | >2000 mg/kg | R41 |

*NC: No Classification

Consequently, it is considered appropriate to adopt a conservative approach and read-across from copper sulphate to copper oxide, recognising that this may result in over-estimation of effects of less bioavailable soluble substances.

Toxicokinetics

• Absorption

Oral absorption of copper has been investigated in human volunteer studies. Absorption of copper occurs primarily in the small intestine. Oral absorption rates have been shown to vary between 12.4 % for subjects with high copper diet and 55.6 % for subjects with a low-copper diet. Absorption rate for subjects with adequate diet for copper is 36 %. Based on these studies, an oral absorption factor of 36 % is used in risk characterisation as a realistic case value of copper oral absorption for humans and 25 % for animals. These values were discussed and agreed at the TMIII08.

Quantitative *in-vitro* measurements of human percutaneous copper absorption have been in the range 0.66 to 5.04% of the applied dose. For the purpose of risk assessment, a percutaneous absorption level of copper of 5% was chosen as the worst case value of copper penetration, under the optimum condition including when emulsifiers were added. This value has been agreed at the TMIII08 and is in line with the EU-VRAR for Copper.

Inhalation absorption of copper was not reported. However, for the purposes of risk assessment, a default inhalation absorption level of copper of 100% will be chosen as the worst case value of copper penetration.

• Distribution

Once absorbed by oral route, copper is bound to albumin and transcuprein and then rapidly transported to the liver where it is incorporated to ceruloplasmin, a transport protein that circulates in the organism and deliver the copper to other organs. The liver is the main organ involved in copper distribution and plays a crucial role in copper homeostasis by regulating its release. It should be however noted that a minor fraction of the absorbed dose can directly be distributed to peripheral organs. In both humans and animals, copper is tightly regulated at a cellular level, involving metallothionein and metallochaperones. These regulating molecules prevent from the accumulation of potentially toxic, free copper ions within the cell. In addition to the liver, the brain is another organ which contains relatively high concentrations of copper.

• Metabolisation

Copper oxide is not exactly metabolised. In specific conditions, it will generate cupric ion which is considered as the actual active substance.

• Elimination

Biliary excretion, with subsequent elimination in the faeces, represents the main route of excretion for copper in animals (rats) and humans, with an excretion rate approximately of 1.7 mgCu/day in humans. Available data show that copper is excreted in the bile in a relatively inabsorbable form. Consequently, little enterohepatic absorption takes place. Biliary excretion of copper and elimination in the faeces is recognised to be essential to the homeostatic regulation of copper in animals and humans

A small amount of copper is also excreted in urine and sweat.

Acute toxicity

Acute toxicity studies were performed with copper oxide in Sprague-Dawley rats by oral and dermal routes. An acute oral study (A6.1.2/01) revealed a $LD_{50} > 2500$ mg/kg bw, not leading to any classification. In dermal acute study (A6.1.2/01) revealed no deaths or clinical signs up to 2000 mg/kg bw. No acute inhalation toxicity study was presented in the dossier, but as far as the Voluntary Risk Assessment (VRA) draft on copper and copper salts recommended to classify all relatively insoluble copper salts as **Xn/R20** (harmful by inhalation), it was agreed with the applicant that such a classification by default could be adopted in order to comply with the VRA recommendation and to minimise animal testing.

Local toxicity

Whereas no sign of skin irritation was reported with copper oxide (A6.1.4/01), a moderate eye irritation was noted through diffuse or translucent corneal opacity, iridal inflammation and moderate conjunctival irritation (A6.1.4/02). However these findings do not meet criteria for a classification. Copper oxide was not considered as sensitising since all the animals tested according to the maximisation test of Magnusson and Kligman displayed negative results (A6.1.5/02).

Repeated-dose toxicity

As mentioned above, chronic, genotoxic, carcinogenic, reprotoxicological and neurotoxicological data were obtained from studies performed in other copper salts (than copper oxide), mainly copper sulphate.

A subchronic study (A6.4.1/01) revealed a nephrotoxic effect, consisting in an increase of cytoplasmic protein droplets in rats when exposed for 92 days through a diet containing copper sulphate. Males and females exhibited this effect at 1000 ppm (corresponding to a dose of 16.3 and 17.3 mgCu/kg bw/day, respectively). This finding was chosen as the critical effect for the risk assessment, by using a NOAEL of 1000 ppm (16.3 mgCu/kg bw/day) for males and females. Other findings such as liver inflammation and lesions of the forestomach were also reported at 2000 ppm and above (corresponding to doses from 34 mgCu/kg bw/day).

Mice equally displayed forestomach lesions when exposed through diet to copper sulphate for 92 days but this occurred at a much higher dose (4000 ppm, corresponding to 187 and 267 mgCu/kg bw/day in males and females, respectively). No other damage was observed in mice (A6.4.1/02).

Mutagenicity

• In vitro tests

There was no evidence of mutagenic activity in *Salmonella typhimurium* strains in the presence or absence of the metabolic activation system when tested with copper sulphate pentahydrate (A6.6.1). Although limited, these *in vitro* data were deemed sufficient and no further *in vitro* assays were required, considering the results of the *in vivo* tests.

• In vivo tests

In vivo studies, conducted with copper sulphate pentahydrate, induced neither micronuclei in the polychromatic erythrocytes from the bone marrow of mice (A6.6.4/01), nor DNA damage in a rat hepatocyte UDS assay (A6.6.4/02).

Equivocal results of additional *in vivo* genotoxicity studies from the public domain (Bhunya and Pati, 1987; Agarwal *et al.*, 1990; Tinwell and Ashby, 1990) but these studies do not meet the higher reliability criteria (1 or 2) under the BPD.

Copper is therefore considered as non genotoxic.

Carcinogenicity

No carcinogenic potential of copper sulphate was detected in rats and mice. However, all available data are of limited value to evaluate the carcinogenic potential of copper compounds. Study durations are in particular too short (<2 years) and group sizes are small for drawing formal conclusions. However, due to the lack of genotoxicity and considering that the expected level of exposure (as described in paragraph 2.2.1.2) is significantly lower than the usual dietary intake of copper (2-3 mg/day), there is no need to conduct new carcinogenicity studies according to OECD guideline 451/453.

Reproductive toxicity

• Developmental toxicity

A developmental study in mice (A6.8.1/01) was submitted but suffers from major methodological deficiencies including no information on maternal toxicity and is neither adequate for classification and labelling nor for risk assessment. No teratogenicity study in the rabbit was submitted, although regulatorily required.

In spite of this lack of (valid) data regarding the developmental toxicity, no further developmental studies were requested for the following reasons:

- Copper is an essential element to fetal development in humans. At birth, the copper level in the newborn baby is about 15 mg, coming from the mother. The copper is absorbed across the placenta and is required for healthy growth and development, especially in blood maturation, bone development, heart development and function of 20 enzymes (Ralph and McArdle, 2001).
- The systemic exposure resulting from primary and secondary to copper is low. As a matter of comparison, it is significantly lower than the usual dietary intake of copper (2-3 mg/day).
- A two-generation oral reproduction study (A6.8.2/06), although not designed for extensively studying malformations, does not raise any particular suspicion of teratogenic potential of copper in the rat. Performed in accordance with OECD test guideline 416, it provides information on the effects of repeated exposure to the substance during all phases of the reproductive cycle including gestation. In particular, the study provides information on the reproductive parameters, and on development, growth and survival of offspring.

On the basis of the above-mentioned data, it is therefore not proposed to classify copper as a teratogenic compound by oral route.

• Fertility

According to the two-generation oral reproduction study (A6.8.2/06), the NOAEL for reproductive toxicity for parental males was 1500 ppm (the highest concentration tested corresponding to 23.6 mg/kg bw/d), The NOAEL for parental females was only 1000 ppm (15.2-35.2 mg/kg bw/d), based on the reduced spleen weight at 1500 ppm. This reduction also occurred in F1 and F2 generations at the same dose level in both males and females. However the reduced spleen weights were not considered a reproductive endpoint as it did not affect growth and fertility.

Therefore as the results of this study do not indicate specific reproductive toxicity at the highest dose level tested, it is proposed that copper sulphate and copper oxide should not be classified as reprotoxic compounds.

Neurotoxicity

From a neurotoxicological point of view, although copper has been recently suspected (Bush and Tanzi, 2008) to be involved in the pathogenesis of the Alzheimer's disease and prion-mediated encephalopathies, this assumption is only speculative and although no valid neurotoxicity study was submitted, no evidence of a neurotoxic potential of copper is suspected from the available studies, in animals or in humans up to now. No further study was therefore deemed necessary.

Human data

Human data are based on both clinical studies and poisoning cases.

Clinical studies were performed in healthy volunteers in order to explore the effect of copper gluconate supplementation in the diet. Even if liver functions were identified as the critical endpoints, no evidence of hepatic damage was reported at 10 mg/day.

Poisoning cases consisted in suicide attempts, food and water contamination (cases of India Childhood Cirrhosis, cases of Childhood Idiopathic Toxicosis) and occupational diseases (through Bordeaux mixture exposure).

Acute symptoms resulted in metallic taste, salivation, epigastric pain, nausea, vomiting and diarrhoea. Anatomo-pathological examinations after self-poisoning (ingestion varying between 1 and 100 g of copper dissolved in water) revealed ulcerations of gastro-intestinal mucosa, hepatic damages (dilatation of central vein, cell necrosis and bile thrombi) and kidney lesions (congestion of glomeruli, swelling or necrosis of tubular cells and sometimes haemoglobin casts).

Chronic symptoms, occurred in a voluntary intoxication by daily ingestion of 30 mg of copper for 2 years and 60 mg during the third year, were malaise, jaundice, hepatomegaly and splenomegaly. Liver examination revealed micronodular cirrhosis. In the particular case of vineyard sprayers intoxication by the Bordeaux mixture (unknown doses), lung lesions with focal distribution were observed: alveoli filled with desquamated macrophages, granuloma in the alveoli septa and fibro-hyaline nodules.

2.2.1.2 EXPOSURE ASSESSMENT

Primary exposure

• <u>Professional exposure</u>

The product in concentrate form is supplied to the timber treatment plants where the timber is impregnated with diluted solutions in enclosed pressure vessels. The treated wood is removed from the vessel at the end of the process and held for drying during a post-treatment conditioning period, before being stored. The major occupational routes of exposure to copper are inhalation and skin contact. Assuming proper hygiene measures are applied, oral exposure would normally not occur in the workplace.

Based on the "Handling, model 1" data from the TNsG on human exposure and following a tiered approach, following exposures were estimated, depending on Hazard Classes:

Table 2.2.1.2-1: Summary of exposure estimates: direct exposure to industrial workers

| Task: Handling of wood and equipment during vacuum-pressure impregnation (including mixing and loading, and post-application) | | | | | | | |
|---|-----------------|----------------------------------|-----------------------------|------------------------------|-----------------------------|-----------------------------|--|
| Users: | Trained indus | strial workers | | | | | |
| Frequency: | 3 cycles/day | 3 cycles/day, 5 hours/day, daily | | | | | |
| | | Inhalation exposure | | Dermal exposure | | Total exposure | |
| Tier - PPE | Hazard Class | Inhaled uptake mg as/day | Systemic dose mg as/kg bw/d | Deposit on skin mg as/day | Systemic dose mg as/kg bw/d | Systemic dose mg as/kg bw/d | |
| Tier 1 : gloves, minimal clothing, | HC 1, 2 or 3 | 0.037 | 0.00061 | 90 | 0.075 | 0.076 | |
| no RPE | HC 4 | 0.068 | 0.00113 | 165 | 0.138 | 0.139 | |
| Tier 2: gloves, | HC 1, 2 or 3 | 0.037 | 0.00061 | 20 | 0.017 | 0.018 | |
| protective clothing, no RPE | HC 4 | 0.068 | 0.00113 | 36 | 0.030 | 0.031 | |

• Non-professional exposure

Intended uses are restricted to professional application, so non-professional primary exposure are not expected.

Secondary exposure

The secondary human exposure assessment considers the potential for the exposure of adults in which they may come into contact with copper treated timber. The scenarios used in this assessment are those contained in the TNsG on Human Exposure part 2 page 77 and User guidance page 51. The following scenarios have been identified as being relevant for assessing the potential exposure of humans to amine copper treated timbers during and after their use:

Adults (workers) - Chronic Handling, cutting and sanding treated timbers

• Adults (consumers) - Acute Handling, cutting and sanding treated timbers

Children - Chronic Playing on preserved timber playground equipment

• Infants - Chronic Playing on preserved timber playground equipment and mouthing of treated timber surface

• Infants - Acute Chewing preserved timber off-cuts

The results of the exposure assessment, depending on the Use Classes are reported in following tables:

Table 2.2.1.2-2: Summary of exposure estimates in secondary human exposure, for HC 4

| Scenario | Route | Estimate * | PPE | Uptake mg as /day | Systemic dose mg as / kg bw |
|--|------------|----------------------------|------------|----------------------|--------------------------------|
| | Inhalation | 1 | None | 0.428 | 0.0071 |
| | Innalation | 1 | RPE | 0.0428 | 0.00071 |
| | | 1 | None | 5.7 | 0.0048 |
| Adulta (nuafassianal) Chuania | D 1 | 1 | Gloves | 0.57 | 0.00048 |
| Adults (professional) - Chronic Handling, cutting and sanding | Dermal | 2 | None | 0.168 | 0.00014 |
| treated timbers | | 2 | Gloves | 0.0168 | 0.000014 |
| | | inhalation 1 | None | 6.128 | 0.0119 |
| | TOTAL | + dermal 1 | RPE+gloves | 0 6128 | 0 00119 |
| | TOTTLE | inhalation 1 | None | 0.596 | 0.0072 |
| | | + dermal 2 | RPE+gloves | 0.0596 | 0.00072 |
| | Inhalation | 1 | None | 0.428 | 0.0071 |
| | Dermal | 1 | None | 5.7 | 0.0048 |
| Adults (consumers) - Acute Handling, cutting and sanding | | 2 | None | 0.168 | 0.00014 |
| treated timbers | TOTAL | inhalation 1 + dermal 1 | None | 6.128 | 0.0119 |
| | | inhalation 1 + dermal 2 | None | 0.596 | 0.0072 |
| Children - Chronic | | 1 | None | 2.72 | 0.0090 |
| Playing on playground structure outdoors | Dermal | 2 | None | 0.080 | 0.00027 |
| | - 1 | 1 | None | 2.72 | 0.014 |
| | Dermal | 2 | None | 0.080 | 0.0004 |
| Infants - Chronic | 01 | 1 | None | 171 | 6.2 |
| Playing on playground structure | Oral | 2 | None | 0.10 | 0.0036 |
| outdoors and mouthing | TOTAL | dermal 1+ oral 1 | None | 173.72 | 6.21 |
| | TOTAL | dermal 2 + oral 2 | None | 0.180 | 0.004 |
| Infants - Acute | Omal | 1 | None | 5.5 | 0.197 |
| Chewing preserved timber off-cuts | Oral | 2 | None | 0.096 | 0.0035 |

^{*:} The different estimates are due to refinements for the estimation of the dislodgeable copper concentration. Estimate $1 = 68\mu g/cm^2$ (unrealistically conservative). Estimate $2 = 2\mu g/cm^2$. The use of PPE (protective gloves and mask) is considered only for professional users.

Table 2.2.1.2-3: Summary of exposure estimates in secondary human exposure, for HC 1-3

| Scenario | Route | Estimate* | PPE | Uptake mg as /day | Systemic dose mg as / kg bw |
|--|------------|----------------------------|--------------|----------------------|--------------------------------|
| | Inholotion | 1 | None | 0.225 | 0.0038 |
| | Inhalation | 1 | RPE | 0.0225 | 0.00038 |
| | | 1 | None | 3 | 0.0025 |
| | Dermal | 1 | Gloves | 0.3 | 0.00025 |
| Adults (professional) - Chronic | Dermai | 2 | None | 0.168 | 0.00014 |
| Handling, cutting and sanding treated timbers | | | Gloves | 0.0168 | 0.000014 |
| treated timbers | | inhalation 1 | None | 3.225 | 0.0063 |
| | TOTAL | + dermal 1 | RPE + gloves | 0.3225 | 0.00063 |
| | TOTAL | inhalation 1 | None | 0.393 | 0.0039 |
| | | + dermal 2 | RPE + gloves | 0.0393 | 0.00039 |
| | Inhalation | 1 | None | 0.225 | 0.0038 |
| A dulta (a angumana) A auta | Dermal | 1 | None | 3 | 0.0025 |
| Adults (consumers) - Acute Handling, cutting and sanding | | 2 | None | 0.168 | 0.00014 |
| treated timbers | TOTAL | inhalation 1 + dermal 1 | None | 3.225 | 0.0063 |
| | | inhalation 1 + dermal 2 | None | 0.393 | 0.0039 |
| Children - Chronic | | 1 | None | 1.43 | 0.0047 |
| Playing on playground structure outdoors | Dermal | 2 | None | 0.080 | 0.00027 |
| | | 1 | None | 1.43 | 0.0074 |
| | Dermal | 2 | None | 0.080 | 0.00027 |
| Infants - Chronic | 0.1 | 1 | None | 90 | 3.26 |
| Playing on playground structure | Oral | 2 | None | 0.10 | 0.0036 |
| outdoors and mouthing | TOTAL | dermal 1 + oral 1 | None | 91.43 | 3.27 |
| | TOTAL | dermal 2 + oral 2 | None | 0.180 | 0.0039 |
| Infants - Acute | Omal | 1 | None | 2.9 | 0.104 |
| Chewing preserved timber off-cuts | Oral | 2 | None | 0.096 | 0.0035 |
| | · | | | | |

^{*:} The different estimates are due to refinements for the estimation of the dislodgeable copper concentration. Estimate $1 = 68\mu g/cm^2$ (unrealistically conservative). Estimate $2 = 2\mu g/cm^2$. The use of PPE (protective gloves and mask) is considered only for professional users.

2.2.1.3 RISK CHARACTERISATION FOR HUMAN HEALTH

The human health risk characterisation is performed using both the AEL and the MOE approaches.

AELs determination

For each exposure scenario, an appropriate AEL is determined on the basis of the exposure frequency. Accordingly, three types of AELs are classically derived: AEL_{short-term}, AEL_{medium-term} and AEL_{long-term} corresponding to short-, medium- and long-term exposures respectively. AELs are usually derived by applying the following formula:

$$AEL = \frac{NOAEL}{Assessment factors}$$

In the case of copper oxide, all AELs (AEL $_{short-term}$, AEL $_{medium-term}$ and AEL $_{long-term}$) were derived on the basis of the NOAEL of 1000 ppm, corresponding to 16.3 mgCu/kg bw/day obtained in the 90-day oral rat study with copper sulphate (A6.4.1/01). An oral absorption rate of 25% was taken into account for calculating the systemic NOAEL as follows:

$$NOAELsystemic = 16.3 \times 0.25 = 4.1 \, mgCu/kg \, bw/d$$

Although copper oxide induced local effects (eye irritation) and copper sulphate induced a minimal to moderate hyperplasia of the squamous mucosa in the forestomach, no local AEC was derived as far as no local effect was detected in the absence of systemic effects. Local effects are therefore covered by systemic AELs.

Regarding the assessment factors, after refinement, a value of 50 (including an inter-species factor of 5 and an intra-species factor of $10)^5$ was applied for deriving $AEL_{short-term}$ and $AEL_{medium-term}$. An additional factor of 2 was integrated for taking into account the duration extrapolation from subchronic to chronic exposures. An overall assessment factor of 100 was therefore adopted for deriving $AEL_{long-term}$.

These values are used as the reference margin of exposure (MOE_{ref}).

The following AELs were therefore derived:

• AEL_{short-term} = 4.1 / 50 = 0.082 mgCu/kg bw/day

• AEL_{medium-term} = 4.1 / 50 = 0.082 mgCu/kg bw/day

• AEL_{long-term} = 4.1 / 100 = 0.041 mgCu/kg bw/day

In the AEL approach, a risk is considered as acceptable if AEL > exposure.

_

⁵ Although the inter-species factor is usually set at 10, it was agreed at TM I09 it could be reduced from 10 to 5 in the case of copper compounds. This factor is composed of an allometric scaling subfactor (which is 4 for rats) and a residual subfactor of 2.5 accounting for the other interspecies variability. Whereas the allometric scaling subfactor was kept unchanged, it was proposed to reduce the residual subfactor from 2.5 to 1.25 on the basis of the extensive toxicokinetic data set in both humans and animals (rats) which demonstrates similarities between the two species in absorption, distribution and excretion of copper compounds.

This approach was accepted by the TCNES and subsequently agreed during the review process by SCHER. The Biocides Technical Meeting adopted it as a refined tier in order to harmonise with the overall assessment factor used in the VRA.

In practice, exposure is expressed as a percentage of the AEL (%AEL). The risk is therefore considered as acceptable if %AEL < 100.

In the MOE approach, a risk is considered as acceptable if $MOE > MOE_{ref}$ (where

$$MOE = \frac{NOAEL}{Exposure})$$

ADI Determination

As no food risk assessment was deemed necessary because of the negligible exposure through food, no ADI was derived.

An ADI value of 0.15 mg/kg bw/d is nevertheless available in the literature (EFSA, 2008).

Risk characterisation for primary exposure scenarios

Professional users

The %AELs and the Margins of Exposure (MOE) were calculated for long-term exposures as reported in the table below:

Table 2.2.1.3-1: Summary of risk assessment for professional users during long-term exposure.

| Task: Handling of wood and equipment during vacuum-pressure impregnation (including mixing and loading, and post-application) | | | | | | | |
|---|------------------|--------------------------------|--------------------------------|---------|-------|----------------|---------|
| Users: | Trained in | ndustrial workers | | | | | |
| | Chronic Toxicity | | | | | | |
| Tier - PPE | Hazard Class | Exposure path | Systemic dose mg as/kg bw/d | Ref MOE | MOE | Ref AEL | Expo as |
| | Class | | ing as/ing b w/a | | | (mgCu/kg bw/d) | % AEL |
| Tier 1: gloves, | HC 1, 2 or 3 | Inhalation and dermal exposure | 0.076 | 100 | 53.9 | 0.041 | 185 |
| minimal clothing, no RPE | HC 4 | Inhalation and dermal exposure | 0.139 | 100 | 29.5 | 0.041 | 339 |
| Tier 2: gloves, | HC 1, 2 or 3 | Inhalation and dermal exposure | 0.018 | 100 | 227.8 | 0.041 | 44 |
| protective clothing, no RPE | HC 4 | Inhalation and dermal exposure | 0.031 | 100 | 132.3 | 0.041 | 76 |

When PPE (protective clothing and new gloves) are worn, it can be seen that the % of AEL and MOE become acceptable whatever the hazard class considered.

In conclusion, since the MOE is greater than the MOE_{ref} and the %AEL is lower than 100%, the risks for industrial users wearing PPE are considered to be acceptable even if chronically exposed.

• Non-professional users

The biocidal product is foreseen to be used by trained professionals only. Thus, a risk characterisation for non-professionals is not relevant.

Secondary exposure as a result of use of the active substance in biocidal product

The %AELs and the Margins of Exposure (MOE) were calculated for secondary exposure scenarios as reported in the tables below:

Table 2.2.1.3-2: Summary of Risk assessment for secondary exposure / Use Class HC 1-3 (Copper loading: 1.80 kg/m³)

| Scenario | Exposure path | Estimate * | Total exposure (mgCu/kg bw/d) | MOE_{ref} | MOE | AEL (mgCu/kg bw/d) | Expo as % AEL |
|--|-----------------------------|------------|----------------------------------|-------------|---------------|-----------------------|------------------|
| Adults (professional) - Chronic Handling, cutting and sanding treated timbers | Inhalation and dermal | 1 + PPE | 0.0063 0.00063 | 100 | 651 6508 | 0.041 | 15.4 1.5 |
| | | 2 + PPE | 0.0039 0.00039 | 100 | 1051 10513 | 0.041 | 9.5 1.0 |
| Adults (consumers) - Acute | Inhalation and | 1 | 0.0063 | 50 | 651 | 0.082 | 7.7 |
| Handling, cutting and sanding treated timbers | dermal | 2 | 0.0039 | 50 | 1051 | 0.082 | 4.8 |
| Children - Chronic Playing on playground structure outdoors | dermal | 1 | 0.0047 | 100 | 872 | 0.041 | 11.5 |
| | | 2 | 0.00027 | 100 | 15185 | 0.041 | 0.7 |
| Infants - Chronic Playing on playground structure outdoors and mouthing | Dermal and | 1 | 3.27 | 100 | 1.25 | 0.041 | 7975 |
| | oral | 2 | 0.0039 | 100 | 1051 | 0.041 | 9.5 |
| Infants - Acute Chewing preserved timber off-cuts | 1 | 1 | 0.104 | 50 | 39.4 | 0.082 | 127 |
| | oral | 2 | 0.0035 | 50 | 1171 | 0.082 | 4.3 |

^{*}The different estimates are due to refinements for the estimation of the dislodgeable copper concentration. Estimate $1 = 68\mu g/cm^2$ (unrealistically conservative). Estimate $2 = 2\mu g/cm^2$. The use of PPE (protective gloves and mask) is considered only for professional users.

The % of AEL on the "Infant playing on playground structure outdoors and mouthing" scenario (Estimate 1) is very largely >100 % (and the MOE is <100). However, it is considered that the exposure value is based on an unrealistic estimate of the dislodgeable copper. By using the more realistic estimate 2, it can be seen that the % of AEL and MOE become acceptable.

Likewise, the % of AEL on the "Infant chewing preserved timber off-cuts" scenario (Estimate 1) is also above 100 % (and the MOE is < 50). However, it is considered that the model in the TNsG Human Exposure is unrealistic as far as it is unlikely that an infant could chew a piece of timber 4cm x 4 cm x 1 cm and certainly would not be able to generate enough saliva to extract wood preservative from the inside of the block of treated wood. It is rather proposed that the infant can remove the dislodgeable residues of copper from the surface of the wood and ingest this material. Treated wood is very hard and is highly likely to be distasteful for the infants. The infant would probably also expel unpleasant tasting materials from his/her mouth. By taking into account these elements in the more realistic estimate 2, it can be seen that the % AEL > 100 and MOE < MOE_{ref}, risk is then considered as acceptable.

For all other *scenarii*, % of AEL and MOE values are acceptable in estimates 1 and 2, the risks for professionals and consumers under the conditions specified above are then acceptable.

Table 2.2.1.3-3: Summary of Risk assessment for secondary exposure / HC 4 (Copper loading: 3.42 kg/m³)

| Scenario | Exposure path | Estimate * | Total exposure mg as / kg bw | MOE_{ref} | MOE | AEL (mgCu/kg bw/d) | Expo as % AEL |
|---|-------------------|------------|------------------------------------|-------------|-------------|-----------------------|------------------|
| Adults (professional) - Chronic Handling, cutting and | Inhalation and | 1 + PPE | 0.0119 0.00119 | 100 | 345 3445 | 0.041 | 29.0 2.9 |
| sanding treated timbers | dermal | 2 + PPE | 0.0072 0.00072 | 100 | 569 5694 | 0.041 | 17.6 1.8 |
| Adults (consumers) - Acute | Inhalation and | 1 | 0.0119 | 50 | 345 | 0.082 | 14.5 |
| Handling, cutting and sanding treated timbers | dermal | 2 | 0.0072 | 50 | 569 | 0.082 | 8.8 |
| Children - Chronic | dermal | 1 | 0.0090 | 100 | 456 | 0.041 | 22.0 |
| Playing on playground structure outdoors | dermar | 2 | 0.00027 | 100 | 15185 | 0.041 | 0.7 |
| Infants - Chronic Playing on playground structure outdoors and mouthing | Dermal and | 1 | 6.21 | 100 | 0.66 | 0.041 | 15146 |
| | oral | 2 | 0.004 | 100 | 1 025 | 0.041 | 9.8 |
| Infants - Acute | oral | 1 | 0.197 | 50 | 21 | 0.082 | 240 |
| Chewing preserved timber off-cuts | | 2 | 0.0035 | 50 | 1171 | 0.082 | 4.3 |

^{*:} The different estimates are due to refinements for the estimation of the dislodgeable copper concentration. Estimate $1=68\mu g/cm^2$ (unrealistically conservative). Estimate $2=2\mu g/cm^2$.. The use of PPE (protective gloves and mask) is considered only for professional users. As previously for class HC 1, 2 and 3 uses, a refinement was necessary for scenarios "Infants playing on playground structure outdoors and mouthing" and "Infant chewing preserved timber off-cuts". Since the estimate 1 was unrealistic, it is proposed to rather consider the more realistic estimate 2 (described above). It has been demonstrated that the risks are acceptable for these estimates 2.

For all other *scenarii*, % of AEL and MOE values are acceptable in estimates 1 and 2, the risks for professionals and consumers under the conditions specified above are then acceptable.

For all other scenarii, % of AEL and MOE values are acceptable in estimates 1 and 2, the risks for professionals and consumers under the conditions specified above are then acceptable.

Overall conclusion of the risk characterisation for human health

The proposed uses of copper oxide can be considered as safe for industrial users, professional users and indirectly exposed persons of the general public.

2.2.2 Environment risk assessment

Copper is applied in wood preservatives in the form of aqueous solutions of copper salts. The environmentally relevant moiety and the active principle of Copper oxide is the cupric ion (Cu²⁺), which may be released to the environment at a low rate.

2.2.2.1 FATE AND DISTRIBUTION IN THE ENVIRONMENT

As a result of the unique fate of copper in water, soil, sediment and sludge, many of the data requirements listed in Section A7 of the Technical notes for Guidance are not applicable for inorganic compounds and metals in particular e.g. hydrolysis, photodegradation and sediment degradation. It is not applicable to discuss copper in terms of degradation half-lives or possible routes of degradation.

Copper oxide as an inorganic compound is not subjected to biological degradation in any environmental compartment. The substance is non-volatile, hydrolytically stable and not biodegradable. Phototransformation in water is not expected. The strong adsorbance to organic carbon, manganese and iron oxides increases in soil with increasing pH.

The most important parameters determining the distribution of copper in the aquatic and soil compartment is adsorption onto solid materials and therefore the copper partitioning coefficients.

Partition coefficient in suspended matter

$$Kpsusp = 30,246 \text{ l/kg (log Kp (pm/w)} = 4.48) (50th percentile)$$

Partition coefficient in sediment

$$Kpsed = 24,409 \text{ l/kg (log Kp(sed/w)} = 4.39) (50th percentile)$$

Partition coefficient in soil

Kpsoil =
$$2 \cdot 120 \cdot 1/\text{kg}$$
 (log Kp (soil/w) = 3.33) (50th percentile)

As all metals, copper becomes complexed to organic and inorganic matter in waters, soil and sediments and this affects copper speciation, bioavailability and toxicity.

Because of the homeostasis of metals, BCF values are not indicative of the potential bioaccumulation. There is therefore limited evidence of accumulation and secondary poisoning of inorganic forms of metals, and biomagnification in food webs.

2.2.2.2 Environmental effect assessment

The risk assessment is carried out on the basis of total concentrations of copper in the environment taking into account the background plus added amount of copper. It was stated that this approach may be more reliable. The PEC values, initially calculated as "added values" were corrected in order to integrate the background concentrations in copper. Total copper concentrations were calculated in taking into account of the natural/pristine or the regional copper background concentrations (as agreed under the Council Regulation (EEC) 793/93 on Existing Substances - EU-RAR).

2.2.2.2.1 Freshwater compartment

For the freshwater pelagic compartment, 139 individual NOEC/EC10 values resulting in 27 different species-specific NOEC values, covering different trophic levels (fish, invertebrates and algae) were used for the PNEC derivation. The large intra-species variabilities in the reported single species NOECs were related to the influence of test media characteristics (e.g., pH, dissolved organic carbon, hardness) on the bioavailability and thus toxicity of copper. Species-specific NOECs were therefore calculated after normalizing the NOECs towards a series of realistic environmental conditions in Europe (typical EU scenario's, with well defined pH, hardness and DOC). Such normalization was done by using chronic copper bioavailability models (Biotic Ligand Models), developed and validated for three taxonomic groups (fish, invertebrates and algae) and additional demonstration of the applicability of the models to a range of other species. The species-specific BLM-normalized NOECs were used for the derivation of log-normal Species Sensitivity Distributions (SSD) and HC5-50 values (the median fifth percentile of the SSD), using statistical extrapolation methods.

The HC5-50 values of the typical EU scenarios ranged between 7.8 to 22.1 µg Cu/L. Additional BLM scenario calculations for a wide range of surface waters across Europe further demonstrated that the HC5-50 of 7.8 µg Cu/L, is protective for 90% of the EU surface waters and can thus be considered as a reasonable worst case for Europe in a generic context.

Copper threshold values were also derived for three high quality mesocosm studies, representing lentic and lotic systems. The mesocosm studies included the assessment of direct and indirect effects to large variety of taxonomic group and integrate potential effects from uptake from water as well as from food.

BLM-calculated HC5-50 values (Assessment Factor (AF)=1) were used as PNEC for the risk characterisation.

The AF=1 was chosen due to the certainty concerning 1) the mechanism of action; 2) the overall evaluation of the database; 3) the robustness of the HC5-50 values; 4) corrections for bioavailability (reducing uncertainty); 5) the sensitivity analysis with regards to DOC and read-across assumptions; 6) the factor of conservatism "built in into" the data and assessment (such as no acclimation of the test organisms and no pre equilibration of test media); 7) results from multi-species mesocosm studies and 8) comparison with natural backgrounds and optimal concentration ranges for copper, an essential metal.

The HC5-50, with an AF=1, was used to derive a PNEC_{freshwater} of 7.8 μ g Cu/l for Europe in a generic context in absence of site-specific information on bioavailability parameters (pH, DOC, hardness).

2.2.2.2. Sediment compartment

The sediment PNEC included using a weight of evidence approach considering different sources and tiered approaches of information: (1) sediment ecotoxicity data, (2) pelagic ecotoxicity data in combination with Kd values derived through different approaches, (3) soil ecotoxicity data and soil bioavailability models and (4) mesocosm/field ecotoxicity.

High-quality chronic benthic NOECs for six benthic species, representing 62 NOEC values were retained for the PNEC derivation. NOEC values were related to sediment characteristics (e.g., Organic Carbon (OC) and Acid Volatile Sulphides (AVS)), influencing the bioavailability and thus toxicity of

copper to benthic organisms. The derivation of the freshwater HC5-50sediment for copper was therefore based on the OC-normalized dataset, containing only low-AVS sediments. Using the lognormal species sensitivity distribution a freshwater HC5-50sediment of 1741 mg Cu/kg OC was derived through the statistical extrapolation method.

Using the equilibrium partitioning (EP) approach, the derived HC5-50sediment (EP) values were comparable or higher than the HC5-50 derived from whole sediment tests. The comparison between the sensitivity of soil and benthic organisms added weight to the HC5-50 from whole sediment tests. The same did sediment threshold values and benthic NOECs that were obtained from four mesocosm studies and one field cohort study.

The AF of 1 was chosen due to the certainty concerning 1) weight of evidence provided; 2) the overall quality of the database; 3) the robustness of the HC5-50 values; 4) corrections for bioavailability (reducing uncertainty); 5) the conservative factor built into the system (no acclimation of the test organisms and only low AVS sediments retained); 6) validations from multi-species mesocosm studies and field studies and 7) comparison with natural backgrounds and optimal concentration ranges.

In case of natural sediments both the amount of AVS and organic carbon present in the sediment has dictated the observed effect levels for copper and were used for the risk characterisation. In absence of AVS data, a default AVS value of $0.77~\mu mol/kg$ dry weight was used. This value corresponded to the 10th percentile of the AVS obtained from a wide Flemish monitoring database and additional AVS data from other European countries.

The HC5-50, with an AF=1, was used to estimate a PNEC_{sediment} of 1741 mg Cu/kg OC, for Europe in a generic context. This corresponding to 87 mg Cu/kg dry weight for a sediment with 5 % O.C.(TGD default value).

2.2.2.2.3 Terrestrial compartment

A high-quality dataset of 252 individual chronic NOEC/EC10 values from 28 different species and processes representing different trophic levels (i.e., decomposers, primary producers, primary consumers) has been retained for the PNEC derivation. The observed intra-species differences in toxicity data were related to differences in bioavailability, the latter related to differences in soil properties and to differences in ageing and application mode and rate.

The soil property best explaining the variability in toxicity for most of the endpoints was the eCEC (effective Cation Exchange Capacity).

For the normalisation of the ecotoxicity data, the respective Cu background concentrations were added on all NOEC/EC10 values which were subsequently normalised to representative EU soils using the relevant regression (bio)availability models, generating soil-type specific HC5-50 values.

Species Sensitivity Distributions were constructed using the normalised NOEC/EC10 data. HC5-50 values from log-normal distributions ranging between 13.2 and 94.4 mg Cu/kg dry weight were obtained. A total of eight single species studies were available in which the toxicity of Cu to microorganisms, invertebrates and plants in field-contaminated aged soils was investigated for a wide range of European soil types (peaty, sandy, clay). A total of five multi-species studies were available, three of which studied the effects of copper in freshly spiked soils and 2 in field contaminated aged soils. Invertebrates, plants and micro-organisms were studied. Single species and multi-species field studies indicate that effects did not occur at an exposure level at the HC5-50-value.

Normalized HC5-50 values (AF=1) were used as PNEC_{soil} for the risk characterisation.

The uncertainty analysis that provides arguments for the AF=1 was based on: 1) the overall quality of the database and the end-points covered; 2) the diversity and representativeness of the taxonomic groups covered by the database; 3) corrections for differences in bioavailability (soil properties); 4) the

statistical uncertainties around the 5th percentile estimate; 5) NOEC values below the HC5-50 and 6) field and mesocosm studies and comparisons of their results with the HC5-50.

To account for the observed difference between lab-spiked soils and field-contaminated soils, a conservative leaching-ageing factor of 2 was agreed based on test data from the mechanistic research on ageing and ionic strength (leaching) effects.

For the PT08 biocidal product dossiers, unlikely to the VRA, a leaching ageing "L/A" factor of 2 was not used to derive the PNECsoil but it was taken into account in the assessment of the PEC soil (PEC divided by 2). Indeed it was stated that decrease of Cu toxicity with ageing has to be taken into account but rather in the exposure assessment than in the hazard assessment. Since this factor was determined over a period of 18 months, it can be applied for PEC calculation over the same amount of time (i.e. TIME 2 only in the PT08). The L/A factor of 2 was used in the PEC (PEC/2), while for VRA, the L/A factor was used in the PNEC (PNEC/2). In the VRA, the NOEC added were first multiplied by the L/A factor (2). The background concentrations from corresponding control soil were then added. All the individual aged NOECtotal were then normalized and finally the HC5-50 was derived; The AF of 1 was applied on the HC5-50..

The HC5-50, with an AF=1, was used to derive a PNEC_{soil} of 45.6 mg Cu/kg dry weight for Europe in absence of site-specific information on soil properties.

2.2.2.2.4 STP compartment

For the STP compartment, high-quality NOECs from respiration or nitrification inhibition studies, relevant to the functioning of a Sewage Treatment Plant (STP), resulted from biodegradation/removal studies and NOECs for ciliated protozoa were used to derive the PNEC for STP micro-organisms.

The lowest reliable observed NOEC value was noted for the inhibition of respiration = 0.23 mg/l expressed as dissolved copper and carried forward as PNEC_{STP} to the risk characterisation.

2.2.2.2.5 Summary of PNECs

| Compartment | PNEC | Unit |
|-------------|-------|--------------|
| STP | 0.23 | [mg.L-1] |
| Freshwater | 7.8 | [µg.L-1] |
| Sediment | 87 | [mg.kgdwt-1] |
| | 18.9 | [mg.kgwwt-1] |
| Soil | 45.6 | [mg.kgdwt-1] |
| | 40.35 | [mg.kgwwt-1] |

2.2.2.3 RISK CHARACTERISATION FOR THE ENVIRONMENT

The intended use of the representative biocidal product is simple vacuum pressure impregnation for use classes 1-4.

No quantitative exposure assessment has been carried out for the life cycle stages "production" and "formulation of the biocidal product".

No exposure assessment has been performed for the life cycle stage "service life of the treated wood" intended for use classes 1 and 2 (indoor use) assuming negligible emissions to the environment.

The concentrations of copper in the environment were estimated following the recommendations given in the currently available Guidance Documents. Stages of the wood life considered for the exposure assessment are industrial wood treatment, storage of treated wood and wood in-service for use class 3 and use class 4.

The emissions to the environment from the stages of storage and wood in-service were calculated on the basis of the results of leaching tests with treated wood in contact with water. Nevertheless, none of the studies submitted by the applicant for the determination of the leaching rates completely comply with the current requirements for this type of test. As a first Tier, calculations are also made assuming a 100% leaching of copper from treated wood over the life service of the vacuum pressure impregnation treatment (20 years).

As cooper is a natural endogenous compound, the releases due to its use as wood preservative have been added to the background environmental concentration. In a first step, the added predicted concentrations of copper were calculated, in line with the equation given by the ESD. In a second step, the added values were corrected in order to integrate the natural/pristine or the regional background concentrations in copper (as agreed under the Council Regulation (EEC) 793/93 on Existing Substances - EU-RAR):

- Natural/pristine background Cu concentrations in water, sediment and soil were taken from the FOREGS Geochemical Baseline Programme (FGBP) database published in March 2004 (http://www.gsf.fi/foregs/geochem/),
- Regional background Cu concentrations in water, sediment and soil were taken from the EU Existing Chemical Regulation.

| Compartment Natural/pristine background concentration | | Regional background concentration | Unit | |
|---|------------|-----------------------------------|--|--|
| Surface water | 0.88 | 2.9 | [μg.L ⁻¹] | |
| Ground water | 0.88 | 2.9 | [µg.L ⁻¹] | |
| Soil | 12 10.6 | 24.4 21.6 | [mg kg _{dwt} -1] [mg kg _{wwt} -1] | |
| Sediment | 21 4.56 | 67.5 14.7 | [mg kg _{dwt} ⁻¹] [mg kg _{wwt} ⁻¹] | |

In the specific case of copper release to soil, the applicant presented studies on copper toxicity in aged contaminated soils. Results from these studies have been reviewed by RMS. They show that, after 18

months ageing, NOECs increased for plants and invertebrates corresponding to a decrease of copper toxicity threshold. For micro-organisms, NOECS increased also but this is also probably due to an adaptation to copper. 18 months ageing tests were however not long enough to show a total remove of toxicity.

The applicant used these data to derive a lab to field factor reflecting the decrease in bioavailability of copper after 18 months, and proposed to apply this factor for the PNEC derivation. Possible underlying mechanisms were detailed and RMS considers that decrease of Cu toxicity with ageing has to be taken into account, but rather in the exposure assessment than in the hazard assessment. Since this factor was determined over a period of 18 months, it can be applied for PEC calculation over the same amount of time or higher (*i.e.*, TIME 2 only).

Therefore, an ageing factor of 2 was applied on the total copper concentrations in soil for the values calculated in TIME 2, in order to consider the phenomenon of copper ageing in soil. This strategy was validated at TMIII08.

2.2.2.3.1 Aquatic compartment (including sediment)

Estimated risks from **vacuum pressure process** indicate an unacceptable risk to surface water and sediment when results are expressed as total concentration, either considering a natural or a regional background concentration. Therefore, during the application process, the product must be re-cycled within the facility or collected and disposed of according to local authority regulations in order to minimise the release to the environment. The RMS proposed to add a label statement on all copper-containing wood preservatives used in vacuum pressure treatment facilities:

"only to be used at timber treatment installations with no surface drains connection to STP in the contained area of the plant"

Estimated risks from **storage** of wood treated by vacuum pressure impregnation indicate an unacceptable risk to the aquatic and terrestrial environment whatever the background considered. All timbers treated by industrial process will have to be stored on impermeable hard standing to prevent direct losses surface water and to allow losses to be collected for disposal.

Concerning wood-in-service, different representative scenarios are considered:

- Use Class 3 Noise barrier (indirect release via STP) and bridge over the pond
- Use Class 4-2 Sheet pilling and Water jetty

These scenarios indicate an acceptable risk for surface water and sediment when releases from wood treated by vacuum pressure impregnation are directed to a sewage treatment plant (noise barrier scenario) but show unacceptable risk when surface treated wood is located above or in a water body (bridge scenario, sheet pilling scenario) even when removal processes are considered (adsorption onto suspended matter and sediment and time-weighted averaging of concentrations). For the scenario water jetty, release to fresh water is acceptable to freshwater compartment when removal processes are taken into account even in considering a worst case situation of 100% leaching rate of applied copper during TIME2 and whatever the background used. But for this same scenario, risks are expected for sediment even in considering a Tier 2 approach and whatever the background used.

Since an unacceptable risk is identified where direct loses to water are possible there should be a labelling against applications where direct losses to water are possible, thereby preventing use in these situations.

2.2.2.3.2 <u>Sewage treatment plant organism</u>

PEC/PNEC ratios indicate acceptable risks to sewage treatment plant either from industrial application by vacuum pressure impregnation, or from wood-in-service releases (noise barrier scenario).

2.2.2.3.3 Atmosphere

From copper-treated wood, there would be no exposure via the atmosphere due to the very low vapour pressure of copper and copper compounds. Therefore, copper-treated wood would not pose an unacceptable risk to the air compartment.

2.2.2.3.4 Terrestrial compartment

The vacuum pressure impregnation application scenario considers the exposure of the soil compartment via the application of contaminated sludge to soil. No risk to terrestrial organisms is expected. However, due to the risk identified for sediment in this scenario (considering a natural or a regional background concentration), releases via STP during industrial application should not be allowed.

The outdoor storage of treated wood on bare soil (class 3) following the industrial vacuum pressure impregnation is expected to pose a risk to soil organisms, even in considering an ageing factor of 2. Storage on bare soil should not be allowed. The emissions from treated wood to soil should be substantially reduced by covering the storage area with protective roof or covering the soil with impermeable coating e.g. concrete. Leachates should be collected and treated appropriately (e.g. incineration).

Industrial treatment for use class 1 and 2 and subsequent covered storage should not cause a risk to the terrestrial organisms. As above, however, due to the risk identified for surface water in these scenarios, releases via STP during industrial application should not be allowed.

Concerning the use class 3 and 4-1, the PEC/PNEC ratios calculated for wood in service treated by industrial vacuum pressure impregnation are below 1 for all the corresponding scenarios (house, fence, noise barrier, fence post and transmission pole) considering a natural or a regional background concentration and a Tier 2 approach for House scenario. For this last scenario, the risk in Time 2 was considered acceptable seeing that PEC/PNEC ratios are equal or only slightly above the trigger value of 1 and that leaching rates from treated wood were probable overestimated. Therefore, these results indicate an acceptable risk for terrestrial organisms. The wood treated by vacuum pressure impregnation should be allowed for a use in class 3 and 4-1.

2.2.2.3.5 Groundwater

The ESD proposes some models to evaluate the amount of compound reaching groundwater by leaching from soil. However, as noticed by the applicant, and agreed by the RMS none of the model proposed (PELMO or PEARL) can be used to estimate groundwater concentrations of a metal like copper.

The fact that no groundwater concentration estimation was presented was justified by the applicant by the fact that copper is immobile in soil.

Nevertheless exposure of groundwater from copper-treated wood was estimated for the house scenario, as a worst case situation of soil contamination, and compared to the reference value for drinking water (2.0 mg/L).

Direct emissions to soil, and subsequently to groundwater, from house are not presenting any unacceptable risk for groundwater compartment. The risk is acceptable even in considering 100% leaching during TIME 2 and whatever the background used (natural or regional). The house scenario is considered to be worst case situation covering all the other hazard class 3 scenarios for treated timber during its service life. Moreover the risk characterisation presented for groundwater was based on the concentration of copper in pore water which is probably an overestimation of the concentration in ground water.

2.2.2.3.6 Non compartment specific effects relevant to the food chain (secondary poisoning)

Copper is an essential micronutrient, needed for optimal growth and development of micro-organisms, plants, animals and humans. Copper acts as an active cofactor in over 20 enzymes and proteins. To ensure appropriate copper tissue levels without causing toxicity from copper excess, internal copper levels are homeostatically regulated by all living organisms. Homeostatic regulation of copper allows organisms, within certain limits, to maintain their total body copper level and to maintain physiologically required levels of copper in their various tissues, both at low and high copper intakes. In the aquatic environment, homeostatic regulation of invertebrates and fish resulted in an inverse relationship between copper BCFs and concentrations in the water. The importance of such homeostasis regulation was recognised in the regulatory framework of aquatic hazard classification (OECD, 2001). Similarly, in terrestrial plants, copper BCFs were inversely related to copper levels in soils.

The molecular mechanism of copper homeostasis, is related to 2 key elements: P-type ATPases that can pump copper across biological membranes in either direction and copper chaperones, important for the intracellular copper homeostasis. This cellular copper homeostasis mechanism is considered as being universal as the sequences of copper chaperones are highly conserved between species.

Besides these active regulation mechanisms, some groups of organisms have developed additional internal regulation mechanism (molecular binding and sequestration) as a strategy to cope against copper excess.

In higher organisms, dietary copper exposure studies in mammals and humans have shown that the intestinal adsorption/ biliary excretion of copper is regulated with varying dietary intakes. Research indeed demonstrated that copper adsorption in humans can vary between 11 and 75 %, depending on the dietary intake. Similarly, mammals and birds can rely on intestinal adsorption and biliary excretion to maintain internal copper levels with large variation in dietary intakes.

Based on the above information, bioaccumulation and biomagnification of copper are considered as not applicable for copper.

2.2.2.4 PBT ASSESSMENT

Being an inorganic compound, the persistence criteria of DT90, field < 1 year and DT50 at 20° C < 6 months that are laid down in paragraph 85 of Annex VI to the Biocides Directive and in the TNsG on Annex I inclusion are not applicable to copper (II) oxide. According to the latter, the degradation triggers do not necessarily apply if the active substance is included in Annex I with regard to areas of use where a long lasting service-life of the treated material is essential and it is scientifically demonstrated that under field conditions there is no unacceptable accumulation in soil (e.g. that the PEC/PNEC < 1 in soil during storage and the service-life of the treated article)

The application as a wood preservative can be considered as such. It was shown above that for storage the PEC/PNEC ratio is above 1. Copper oxide meet the criteria for persistence in soil, and can be therefore considered as persistent.

Due to the homeostatic regulation process of invertebrates and fish of copper, bioaccumulation and biomagnification of cooper are considered as not applicable for copper.

Considering the HC5-50 value of 7.8 μ g/L for the aquatic compartment, Copper oxide fulfil the T criterion. Copper oxide does not fulfil the PBT-criteria. Therefore inclusion in Annex I is not restricted by these criteria.

2.2.3 Overall summary

| SCENARIO | | Human primary exposure | | Human secondary exposure | | Aquatic | CITID | Terrestrial | Ground | | Secondary | |
|-----------------|---|------------------------|---------------------|--------------------------|----------------|----------------|------------|----------------|------------|-----|-----------|----|
| | | Professional | Non professional | Worker | Consumer | compartment | STP | compartment | water | Air | poisoning | |
| INDUSTRIAL APPI | LICATION SIMPL | E VACUUM PRE | SSURE IMPREG | NATION | | | | | | | | |
| Application | | Acceptable* | NR | Aggentable | Aggantable | Not acceptable | Acceptable | Acceptable | NR | NR | NR | |
| Storage | | Acceptable. | NK | Acceptable | Acceptable | Not acceptable | NR | Not acceptable | NR | NR | NR | |
| Wood in-service | Classes 1-2 | | | | | NR | NR | NR | NR | NR | NR | |
| | Class 3 without direct release to water | | | Acceptable | | | Acceptable | Acceptable | Acceptable | NR | NR | NR |
| | Class 3 with direct release to water | NR | NR | | ble Acceptable | Not acceptable | NR | NR | NR | NR | NR | |
| | Class 4-1 | | | | | | NR | NR | Acceptable | NR | NR | NR |
| | Class 4-2 | | | | | Not acceptable | NR | NR | NR | NR | NR | |

Overall conclusions: Simple vacuum pressure impregnation process should be allowed due to acceptable risk for uses in classes 1 to 4-1 with the following mitigation measures:

- During industrial treatments, collective protective equipment shall be ensured when appropriate, and the operators must wear the appropriate personal protective equipments.
- During industrial application the emissions to surface water have to be forbidden. Appropriate mitigation measures such as waste recycling or incineration have to be performed.
- All timbers treated by industrial process will have to be stored on impermeable hard standing or under a protective roof to prevent direct losses to soil and surface water and to allow losses to be collected and treated appropriately (e.g. incineration).
- Pre-treated timber must not be in contact with or above surface water (a use in class 4-2 is therefore not allowed unless data is submitted to demonstrate that the product will meet the requirements of Article 5 and Annex VI, if necessary by the application of appropriate mitigation measures).

NR: Non relevant

^{*} Considering the wearing of PPE

3 – PROPOSAL FOR THE DECISION

3.1 Background to the proposed decision

The Copper oxide dossier has been submitted by the Wood Preservative Copper Task Force for its use as a wood preservative. The reference biocidal products that will contain copper oxide are limited to a professional use for treating timbers.

The applicant is not currently placing nano forms of copper oxide on the market. Therefore, the submitted dossier and the finalised assessment report don't cover potential nanoforms of this copper compound, should such forms exist.

The evaluation of the dossier led to the following conclusions:

- \Rightarrow The substance is correctly identified with a purity > 97.6% w/w as copper oxide equivalent to 78.0 % copper (w/w).
- There is no concern regarding the physico-chemical properties.
- The submitted information on copper salts proves a sufficient efficacy against wood destroying fungi and insects (including termites).
- The health effects of copper oxide are well documented for acute and irritation endpoints. In contrast, all the chronic data are based on other copper salts. However, considering that copper oxide is a precursor releasing cupric ion, which is the actual active element, it was accepted to base the risk assessment on the provided studies. Toxicological profile in the light of exposure resulting from the use of this substance is of no concern. The conclusion of the toxicological assessment is that the risk for professional users is negligible when safe operational procedures are established and appropriate PPE are used. Risks to indirectly exposed persons (including infants) is also negligible.
- With regard to environmental exposure and effects, based on the risk assessments conducted, it is considered that safe use(s) can only be identified if the possibility of direct surface water exposure is excluded (with the exception of the noise barrier scenario which leads to indirect releases). It is recommended that this should be a condition of Annex I inclusion. The environmental risk assessment indicates that industrial application result in unacceptable risk for the aquatic compartment (application for class 1, 2, 3 and 4) during the application and in unacceptable risk for the aquatic and terrestrial environment during storage (class 3 and 4). The environmental risk assessment for the in-use phase of the treated wood used outdoor (class 3 and 4-1) indicates that the risk to the terrestrial environment is acceptable for all the relevant scenarios. Due to the risk for the aquatic compartment, the use of copper in hazard class 4-2 should not be allowed unless data is submitted to demonstrate that the product will meet the requirements of Article 5 and Annex VI, if necessary by the application of appropriate mitigation measures.
- Classification and labelling as Xn-N (R20, R22, R50-53) is proposed for copper oxide, given the acute oral study in rats and the Voluntary Risk Assessment recommendations on labelling of insoluble copper salts.

22/09/2011

3.2 Proposed decision regarding the inclusion in Annex I or IA

It is proposed that copper (II) oxide (CAS-No. 1317-38-0) be included in Annex I of Council Directive 98/8/EC as an active substance in wood preservative products (product type 8), subject to the following specific provisions:

- 1. The active substance copper oxide shall have a minimum purity ≥ 97.6% w/w equivalent to 78.0 % w/w copper
- 2. The identity and the maximum content of impurities have to comply with the confidential part of the dossier.
- 3. In view of the risk assessment, products shall be authorised only for industrial use by simple vacuum pressure impregnation, unless it is demonstrated in the application for product authorisation that risks to human and environment demonstrate acceptable levels for other types of application in accordance with Article 5 and Annex VI.
- 4. In view of the assumptions made during the risk assessment, products authorised for industrial use must be applied with appropriate personal protective equipment and safe operational procedures should also be established, unless it can be demonstrated in the application for product authorisation that risks to industrial users can be reduced to an acceptable level by other means
- 5. In view of the risks identified for the aquatic and soil compartments, appropriate risk mitigation measures must be taken to protect those compartments. In particular, labels and/ or safety data sheets of products authorised for industrial use shall indicate that freshly treated timber must be stored after treatment under shelter or on impermeable hard standing to prevent direct losses to soil or water and that any losses must be collected for reuse or disposal.
- 6. In view of the risks identified for the aquatic compartments, products shall not be authorised for treatment of wood that will be used in outdoor constructions near or above water, and for the treatment of wood in direct surface water contact, unless data is submitted to demonstrate that the product will meet the requirements of Article 5 and Annex VI, if necessary by the application of appropriate mitigation measures.

Specific provisions n°3 and n°4 are necessary due to risks identified for industrial users (see section 2.2.1.3).

Specific provision n°5 is necessary due to risks identified for the aquatic and soil compartments, during the industrial treatment phase (see section 2.2.2.3.1 and 2.2.2.3.4).

Specific provision 6 is necessary due to risks identified for the aquatic compartment, during service-life of wood continuously exposed to water (see section 2.2.2.3.1).

3.3 Factor to be taken into account by Member States when authorising products

1. The assessment was made for products containing copper oxide only for industrial treatments by simple vacuum pressure impregnation, with negligible inhalation exposure, for uses in hazard classes 1 to 4-1. Other uses will have to be considered at the product authorisation stage.

22/09/2011

- 2. The efficacy data are based on the current state of the art concerning the practical uses of copper-based products during the last decades, related to the expected service life of wooden elements treated with copper-based product. As copper based wood preservative is used in conjunction with other biocides, full efficacy data should be provided by the individual applicants at product authorisation stage including data performed with copper oxide.
- 3. No repeated-dose inhalation study was submitted due to the negligible exposure by this route. A negligible inhalation exposure must therefore be checked when authorising products. Otherwise adequate inhalation data will have to be submitted.
- 4. Concerning environmental risk assessment, Biotic Ligand Modelling (BLM)-calculated HC5-50 values were used as PNEC for risk characterization. Under product authorization, Member States will have to investigate if the BLMs can be applied under their specific conditions. Moreover, the national natural background should be compared to the background level used for the risk assessment.

3.4 Demand for further information

The submitted dossier satisfies the requirements of Annexes IIA, IIIA, IIB and IIIB of Directive 98/8/CE and is sufficient to recommend Annex I inclusion for copper oxide as wood preservative.

Nevertheless, some information or studies, that are not expected to change the conclusion on the inclusion into Annex I, are needed as confirmatory data. They should preferably be submitted to the original Rapporteur Member State (France) at the latest 6 months before the date of inclusion of the active substance into Annex I:

- A new 5-batch analysis is required to support each source and to ensure compliance of current
 production of each source of copper oxide to the specifications. This must include the analysis
 of all relevant and significant impurities using validated methods of analysis. As Nickel has
 been found in some sources of copper coumpounds used as active substances for PT08, nickel
 should be searched in these analysis.
 - If Nickel is present, but its form unknown, it should not be present above its worst case classification limit of 0.01%. Besides, a method of analysis of Nickel should therefore be provided as well.
- For the source of Peninsula Copper Industies Inc., which is not accepted for the moment, further data have to be provided (5 batch analysis, and complementary data about one additional impurity compared to the reference specifications), in order to prove the technical equivalence with the reference specifications.
- The process of manufacture of the source of Adchem should be provided.
- Validation data must be provided for the determination of copper in copper (II) oxide.
- A method for one impurity > 0.1% was not provided and must be provided (see detail in the confidential part of the CAR).
- A validated analytical method for the determination of copper in the biocidal product must be provided.

| Copper | (II) | oxide |
|--------|------|-------|
| COPPE | (/ | |

22/09/2011

• If a demand of autrhorization is submitted for the reference product of the annex I dossier, surface tension and compatibility with other products must be provided.

Appendix 1 – Listing of endpoints

Identity, Physical and Chemical Properties, Details of Uses, Further Chapter 1: Information, and Proposed Classification and Labelling

Active substance (ISO Common Name)

Copper (II) oxide

Function (e.g. fungicide)

Fungicide and insecticide (including termiticide) for

wood preservatives.

Rapporteur Member State

France

Identity (Annex IIA, point II.)

Chemical name (IUPAC)

Chemical name (CA)

CAS No

EC No

Other substance No.

Minimum purity of the active substance as manufactured (g/kg or g/l)

Identity of relevant impurities and additives (substances of concern) in the active substance as

manufactured (g/kg)

Copper (II) oxide

Cupric oxide

1317-38-0

215-269-1

The CIPAC code number for copper compounds is 44.

780 g/kg as copper, equivalent to 976 g/kg copper oxide

Lead: max 0.02%

Cadmium: max 0.0002% Arsenic: max 0.007%

Nickel: max 0.01%

(see detail in the confidential part of the CAR)

Molecular formula

Molecular mass

Structural formula

CuO

79.55 g/mol

Cu=O

22/09/2011

Physical and chemical properties (Annex IIA, point III., unless otherwise indicated)

| Melting point (state purity) | 1446°C (unspecified purity) |
|---|---|
| | (1 1 2) |
| Boiling point (state purity) | Not determined as melting point > 1446 °C |
| Temperature of decomposition | Not determined as melting point > 1446 °C |
| Appearance (state purity) | Dark grey powder at 25 °C |
| | Purity: ≥ 97.8% w/w |
| Relative density (state purity) | 1.018 |
| | Purity: ≥ 97.8% w/w |
| Surface tension | Not required for substances with water solubility <1mg/L. |
| Vapour pressure (in Pa, state temperature) | Not measurable |
| Henry's law constant (Pa m ³ mol ⁻¹) | Not applicable |
| Solubility in water (g/l or mg/l, state temperature) | pH 5.1 – 5.5 (buffered solution): >0.23 g/l at 20°C +/-0.5°C (dependent on acid availability; pH increased during solubilisation). |
| | pH 6 (purified water): 3.94 x 10 ⁻⁴ g/l at 20°C +/- 0.5°C |
| | pH 9 (buffered solution): < 1.0 x 10 ⁻⁵ g/l at 20°C +/- 0.5°C |
| Solubility in organic solvents (in g/l or mg/l, state temperature) (Annex IIIA, point III.1) | Not determined, as copper oxide will be used only in aqueous formulations. |
| Stability in organic solvents used in biocidal products including relevant breakdown products (IIIA, point III.2) | Not determined, as copper oxide will be used only in aqueous formulations. |
| Partition coefficient (log $P_{\rm OW}$) (state temperature) | Due to the nature of the active substance (inorganic salt), log Pow is not a relevant information. Partition coefficients in water, sediments and soils are used instead. |
| Hydrolytic stability (DT ₅₀) (state pH and temperature) (point VII.7.6.2.1) | Not applicable |
| Dissociation constant (not stated in Annex IIA or IIIA; additional data requirement from TNsG) | Cannot be determined. |
| UV/VIS absorption (max.) (if absorption > 290 nm | Molar absorption coefficient (dm³ mol⁻¹.cm⁻¹): 170 |
| state ε at wavelength) | Medium: Acidic (pH 1.2) |
| | Wavelength: 242 nm |
| Photostability (DT_{50}) (aqueous, sunlight, state pH) (point VII.7.6.2.2) | Not applicable. |
| Quantum yield of direct phototransformation in water at $\Sigma > 290$ nm (point VII.7.6.2.2) | Not applicable. |
| Flammability | Not highly flammable. |
| Explosive properties | Not explosive. |
| | |

$\textbf{Classification and proposed labelling} \ (Annex \ IIA, \ point \ IX.)$

| | Directive 67/548/EEC | Regulation 1272/2008 | |
|--|--|--|--|
| with regard to physical/chemical data | Not classified. | | |
| with regard to toxicological data | Xn, Harmful R20, Harmful by inhalation | Acute Tox. 4 /H332 - Harmful if inhaled | |
| with regard to fate and behaviour data | N, dangerous for environment R53, May cause long-term adverse effects in the aquatic environment | Aquatic chronic/H410 - Very toxic to aquatic organisms and may cause long lasting effects in the aquatic environment | |
| with regard to ecotoxicological data | R50, Very toxic to aquatic organisms | Aquatic chronic/H410 - Very toxic to aquatic organisms and may cause long lasting effects in the aquatic environment | |

Chapter 2: Methods of Analysis

Analytical methods for the active substance

Technical active substance (principle of method) (Annex IIA, point 4.1)

Purity is not directly determined but calculated from total copper content. This is possible because other copper forms (i.e. metallic and cuprous) are not expected to be present in the technical materials.

Total copper content can be determined by various well-known methods such as volumetric thiosulphate method (CIPAC E Copper 44/TC/M/3.2), electrogravimetric method (CIPAC E Copper 44/TC/M/3.1). A validation study is required.

Impurities in technical active substance (principle of method) (Annex IIA, point 4.1)

Trace metals, including those of toxicological significance (arsenic, cadmium) can be determined by AAS. Before analysis, the sample is dissolved in an acid mixture and placed on a hotplate until digestion is complete. The AAS methods used to obtain five batch analysis data of impurities in copper oxide are variations on internationally accepted guidelines such as ASTM E53-98 and US EPA methods 206.2, 213.1 and 239.1 for arsenic, cadmium and lead, respectively.

Other suitable methods include Inductively Coupled plasma – Atomic Absorption Spectroscopy (ICP-AES) (e.g. US EPA method 200.7), which is applicable to the determination of µg/l concentrations of a large number of elements in a variety of matrices. Prior to analysis, samples must be solublised or digested using an appropriate method. Samples are nebulised and the resulting aerosol is transported to the plasma torch. Element-specific emission spectra are produced by a fadio-frequency inductively coupled plasma. The spectra are dispersed by a grating spectrometer, and the intensities of the emission lines are monitored by photosensitive devices.

Method for or one impurity > 0.1% and for nickel must be provided

Analytical methods for residues

Soil (principle of method and LOQ) (Annex IIA, point 4.2)

Air (principle of method and LOQ) (Annex IIA, point 4.2)

Water (principle of method and LOQ) (Annex IIA,

ICP-AES methods (e.g. AOAC official method 990.8). The estimated instrumental limit of detection (LOD) is 6 μ g Cu/l (LOQ not determined). Another suitable method is AAS (e.g. US EPA method 7210), with an LOD of 20 μ g Cu/l and a LOQ of 200 μ g Cu/l. For both methods of analysis, the sample must first be digested.

Residues of copper may be determined in air using Flame-AAS or ICP-AES methods (e.g. NIOSH methods 7029 or 7300 respectively). The estimated instrumental limits of determination (LOD) are 0.05 and 0.07 μg Cu/filter (LOQ) not determined

In water, trace elements may be determined by Inductively Coupled Plasma – Mass Spectroscopy (ICP-

point 4.2) MS) (e.g. US EPA method 200.7). The estimated LOQ for this method is 20 µg Cu/l. Other suitable methods include AAS with direct aspiration (LOQ 0.2 mg/l) (e.g. US EPA method 220.1) and AAS with graphite furnace (LOQ 5.0 μ g/l) (e.g. US EPA method 220.2). For all three methods of analysis, the sample must first be digested. ICP-AES may also be used for analysing elements in Body fluids and tissues (principle of method and LOQ) (Annex IIA, point 4.2) body fluids and tissues following acid digestion of the sample. LOQs are $10 \mu g/100 g$ blood, $2 \mu g/g$ tissue (e.g. NIOSH method 8005) and 0.25 μ /sample of urine (NIOSH method 8310). Food/feed of plant origin (principle of method and Not applicable LOQ for methods for monitoring purposes) (Annex IIIA, point IV.1) Food/feed of animal origin (principle of method Not applicable and LOQ for methods for monitoring purposes) (Annex IIIA, point IV.1)

22/09/2011

Copper (II) oxide

Chapter 3: Impact on Human Health

Absorption, distribution, metabolism and excretion in mammals (Annex IIA, point 6.2)

| Rate and extent of oral absorption: | It was agreed during the TMIII09 that an oral absorption of 36% for humans and 25% for animals have to be used. |
|--|--|
| Rate and extent of dermal absorption: | It was agreed during the TMIII09 that a dermal absorption of 5% has to be used for diluted solutions and 100% for the concentrated product. |
| Distribution: | Once absorbed by oral route, copper is bound to albumin and transcuprein and then rapidly transported to the liver where it is incorporated to ceruloplasmin, a transport protein that circulates in the organism and deliver the copper to other organs. The liver is the main organ involved in copper distribution and plays a crucial role in copper homeostasis by regulating its release. It should be however noted that a minor fraction of the absorbed dose can directly be distributed to peripheral organs. In both humans and animals, copper is tightly regulated at a cellular level, involving metallothionein and metallochaperones. These regulating molecules prevent from the accumulation of potentially toxic, free copper ions within the cell. In addition to the liver, the brain is another organ which contains relatively high concentrations of copper. |
| Potential for accumulation: | All mammals have metabolic mechanisms that maintain homeostasis (a balance between metabolic requirements and prevention against toxic accumulation). Because of this regulation of body copper, indices of copper status remain stable except under extreme dietary conditions. This stability was demonstrated in a study in which human volunteers received a diet containing total copper in the range 0.8 to 7.5 mg/d. Under these conditions, there were no significant changes in commonly used indices of copper status, including plasma copper, ceruloplasmin, erythrocyte superoxide dismutase and urinary copper. |
| Rate and extent of excretion: | Biliary excretion is quantitatively the most important route, with a mean copper excretion estimated to be in the order of 1.7 mg Cu/day ($24.6 \pm 12.8 \mu g$ Cu/kg bodyweight). A small amount of copper is also lost in urine and in sweat. Excretion of endogenous copper is influenced by dietary copper intake. When the copper intake is low, turnover is slow and little endogenous copper is excreted and vice versa. Faecal copper losses reflect dietary copper intake with some delay as intake changes and copper balance is achieved. Urinary losses do not contribute to the regulation of copper stores and contribute very little to the overall balance. |
| Toxicologically significant metabolite | None. |
| Acute toxicity (Annex IIA, point 6.1) | |
| Rat LD ₅₀ oral | >2500 mg/kg |
| Rat LD ₅₀ dermal | >2000 mg/kg |
| 50 44 | g g |

Copper oxide

Rat LC₅₀ inhalation

No data (but nevertheless classified as Xn/R20 for harmonising with classification of other insoluble copper salts)

Skin irritation

Negative; not classified as irritating to skin.

Eye irritation

Moderately irritating; not classified as an eye irritant.

Skin sensitization (test method used and result)

Negative; not classified as a skin sensitiser.

Repeated dose toxicity (Annex IIA, point 6.3)

Species/ target / critical effect

The test substance used the following study was copper (II) sulphate.

Rat/ liver/ inflammation

Rat/ kidney/ cytoplasmic droplets

Rat, mouse/ forestomach/ minimal to moderate

hyperplasia of the squamous mucosa

16.3 mgCu/kg bw/day

Not available.

Not available.

Lowest relevant oral NOAEL

Lowest relevant inhalation NOAEL / LOAEL

Lowest relevant dermal NOAEL / LOAEL

Genotoxicity (Annex IIA, point 6.6)

The test substance used in each of the following studies was copper (II) sulphate pentahydrate.

- 1. Ames test in *Salmonella typhimurium* negative in both the presence and absence of S9 mix.
- 2. Bone marrow micronucleus study in the mouse negative at a dose of 447 mg/kg bw.
- 3. *In vivo/in vitro* unscheduled DNA synthesis study in the livers of orally dosed male rats negative, following treatment with doses of 632.5 or 2000 mg/kg bw.

These studies demonstrate that copper is not mutagenic in the *in vitro* and *in vivo* test systems used.

Carcinogenicity (Annex IIA, point 6.4)

Species/type of tumour

Available studies of the carcinogenicity of copper compounds in rats and mice, although not fully reliable, have given no indication that copper salts are carcinogenic.

Not applicable.

lowest dose with tumours

Reproductive toxicity (Annex IIA, point 6.8)

Species/ Reproduction target / critical effect

The test substance used in the following study was copper (II) sulphate pentahydrate.

Rat/Two-generation study/no reproductive effects observed, decreased spleen weight

Lowest relevant reproductive NOAEL / LOAEL

Copper sulphate cannot be regarded as having adverse effects on fertility in the animals tested.

1500 ppm NOAEL in rat two-generation study = 23.6-

Copper oxide

Species/Developmental target / critical effect

Lowest relevant developmental NOAEL / LOAEL

43.8 mgCu/kg bw/d (maximal dose tested)

Mouse/ Developmental toxicity/ malformations (study with major methodological deficiencies)

6 mg Cu/kg bw/d

(NOAEL maternal toxicity = 6 mg Cu/kg bw/d)

However rat two-generation study with copper sulphate pentahydrate does not raise any particular teratogenic concern.

Neurotoxicity / Delayed neurotoxicity (Annex IIIA, point VI.1)

Species/ target/critical effect

Lowest relevant developmental NOAEL / LOAEL.

No evidence for neurotoxic potential from other studies.

Other toxicological studies (Annex IIIA, VI/XI)

| None. | | | |
|-------|--|--|--|
| | | | |

Medical data (Annex IIA, point 6.9)

Direct observation, eg clinical cases, poisoning incidents if available; data point 6.12.2.

Acute symptoms resulted in metallic taste, salivation, epigastric pain, nausea, vomiting and diarrhoea. Anatomo-pathological examinations after self-poisoning (ingestion varying between 1 and 100 g of copper dissolved in water) revealed ulcerations of gastro-intestinal mucosa, hepatic damages (dilatation of central vein, cell necrosis and bile thrombi) and kidney lesions (congestion of glomeruli, swelling or necrosis of tubular cells and sometimes haemoglobin casts).

Chronic symptoms, occurred in a voluntary intoxication by daily ingestion of 30 mg of copper for 2 years and 60 mg during the third year, were malaise, jaundice, hepatomegaly and splenomegaly. Liver examination revealed micronodular cirrhosis. In the particular case of vineyard sprayers intoxication by the Bordeaux mixture (unknown doses), lung lesions with focal distribution were observed: alveoli filled with desquamated macrophages, granuloma in the alveoli septa and fibrohyaline nodules.

Summary (Annex IIA, point 6.10)

ADI

AEL short- and medium-term

AEL long-term

| Value | Study | Safety factor |
|----------------------|-------------|-----------------|
| 0.15 mg/kg bw/day | EFSA (2008) | Not applicable. |
| 0.082 mg/kg bw/d | 90d in rats | MOE ref = 50 |
| 0.041 mg/kg bw/d | 90d in rats | MOE ref = 100 |

| Copper oxide | | |
|--------------|--|--|
| | | |

| Drinking water limit | No data reported |
|-----------------------------|------------------|
| ARfD (acute reference dose) | Not applicable |

Acceptable exposure scenarios

Professional users

Non-professional users

Indirect exposure as a result of use

| Handling of wood and equipment during vacuum- |
|--|
| pressure impregnation |
| (including mixing and loading, and post-application) |

| Users | Trained industrial workers | | | |
|----------------------------|----------------------------|--------------------------------------|--------------------------------------|--|
| Tier - PPE | Hazard Class | Exposure path | Systemic dose mg as/kg bw/d | |
| Tier 2: gloves, protective | HC 1, 2 or 3 | Inhalation and dermal exposure | 0.018 | |
| clothing, no RPE | HC 4 | Inhalation and dermal exposure | 0.031 | |

The biocidal product is foreseen to be used by trained professionals only. Thus, a risk characterisation for non-professionals is not relevant.

Secondary exposure / Use Class HC 1-3 (Copper loading: 1.80 kg/m³)

| ` | 0 0 | | | | |
|-------------------------------------|--------------------------|-------------|--|--|--|
| Scenario | Exposure path | Estimate | Total exposure (mgCu/kg bw/d) | | |
| | Adults (pr | ofessional) | | | |
| Chronic Handling, cutting and | Inhalation | 1 + PPE | 0.0063 0.00063 | | |
| sanding treated timbers | and dermal | 2 + PPE | 0.0039 0.00039 | | |
| Adults (consumers) | | | | | |
| Acute Handling, cutting and sanding | Inhalation and dermal | 1 | 0.0063 | | |
| treated timbers | | 2 | 0.0039 | | |
| Children | | | | | |
| Chronic Playing on playground | Dermal | 1 | 0.0047 | | |
| structure outdoors | Dennai | 2 | 0.00027 | | |
| Infants | | | | | |

| | | | T |
|---|------------------------------------|-------------|---------------------------------------|
| Chronic Playing on playground structure outdoors and mouthing | Dermal and oral | 2 | 0.0039 |
| Acute Chewing preserved timber off- cuts | Oral | 2 | 0.0035 |
| | xposure / Use (ling: 3.42 kg/n | | |
| Scenario | Exposure path | Estimate | Total exposure mg as / kg bw |
| | Adults (pr | ofessional) | |
| Chronic | | 1 | 0.0119 |
| Handling, | Inhalation and dermal | + PPE | 0.00119 |
| cutting and sanding | | 2 | 0.0072 |
| treated timbers | | + PPE | 0.00072 |
| timoers | Adults (co | onsumers) | |
| Acute | | | |
| Handling, cutting and | Inhalation | 1 | 0.0119 |
| sanding treated timbers | and dermal | 2 | 0.0072 |
| | Chil | dren | |
| Chronic Playing on | dermal | 1 | 0.0090 |
| playground structure outdoors | dermai | 2 | 0.00027 |
| | Infa | ants | |
| Chronic Playing on playground structure outdoors and mouthing | Dermal and oral | 2 | 0.004 |
| Acute Chewing preserved timber off- | oral | 2 | 0.0035 |

cuts

Chapter 4: Fate and Behaviour in the Environment

Route and rate of degradation in water (Annex IIA, point 7.6, IIIA, point XII.2.1, 2.2)

Hydrolysis of active substance and relevant metabolites (DT₅₀) (state pH and temperature)

Photolytic / photo-oxidative degradation of active substance and resulting relevant metabolites

Readily biodegradable (yes/no)

Biodegradation in seawater

Non-extractable residues

Distribution in water / sediment systems (active substance)

Not applicable to metals.

The distribution of metals between aqueous phase and soil/sediment/suspended matter should preferentially be described on the basis of measured soil/water, sediment/water and suspended matter/water equilibrium distribution coefficient (TECHNICAL GUIDANCE DOCUMENT on Risk Assessment Part II Appendix VIII, 2003; TECHNICAL GUIDANCE DOCUMENT Annex 4-VIII Environmental risk assessment for metals and metal compounds (RIP 3.2-2).

From the literature overview, the following partitioning coefficients have thus been derived for Cu metal and Cu compounds:

Partition coefficient in suspended matter

 $Kp_{susp} = 30,246 \text{ l/kg (log Kp (pm/w)} = 4.48) (50^{th} \text{ percentile) (Heijerick et al, 2005)}$

Partition coefficient in sediment

Kpsed = 24,409 l/kg (log Kp(sed/w) = 4.39) (50th percentile) (Heijerick *et al.*, 2005)

Not applicable to metals.

Distribution in water / sediment systems (metabolites)

Route and rate of degradation in soil (Annex IIIA, point VII.4, XII.1.1, XII.1.4; Annex VI, para. 85)

Mineralization (aerobic)

Laboratory studies (range or median, with number of measurements, with regression coefficient)

Not relevant for the nature of the active substance which is an inorganic metal salt.

 DT_{50lab} (20°C, aerobic): Not applicable to metals.

DT_{90lab} (20°C, aerobic): Not applicable to metals.

DT_{50lab} (10°C, aerobic): Not applicable to metals.

DT_{50lab} (20°C, anaerobic): Not applicable to metals.

Not applicable to metals.

Field studies (state location, range or median with number of measurements)

DT_{50f}: Not applicable to metals.

DT_{90f}: Not applicable to metals.

Copper oxide

Anaerobic degradation

Soil photolysis

Non-extractable residues

Relevant metabolites - name and/or code, % of applied a.i. (range and maximum)

Soil accumulation and plateau concentration

Not applicable to metals.

Not applicable to metals.

Not applicable to metals.

Not applicable to metals.

Although unable to degrade, the affect of ageing on the distribution of copper in soil results in increased immobilisation by long term adsorption and complexation reactions in the soil.

Adsorption/desorption (Annex IIA, point XII.7.7; Annex IIIA, point XII.1.2)

Ka, Kd

 Ka_{oc} , Kd_{oc}

pH dependence (yes / no) (if yes type of dependence)

The distribution of metals between aqueous phase and soil/sediment/suspended matter should preferentially be described on the basis of measured soil/water, sediment/water and suspended matter/water equilibrium distribution coefficient (TECHNICAL GUIDANCE DOCUMENT on Risk Assessment Part II Appendix VIII, 2003; TECHNICAL GUIDANCE DOCUMENT Annex 4-VIII Environmental risk assessment for metals and metal compounds (RIP 3.2-2).

From the literature overview, the following partitioning coefficients have thus been derived for Cu metal and Cu compounds:

Partition coefficient in soil

 $Kd = 2120 \text{ l/kg (log } K_p = 3.33) (50^{th} \text{ percentile)}$ (Sauvé *et al.* 2000)

| Copper oxide | |
|--------------|--|
|--------------|--|

Fate and behaviour in air (Annex IIIA, point VII.3, VII.5)

| Direct photolysis in air | Not relevant for metals. |
|------------------------------------|--------------------------|
| Quantum yield of direct photolysis | Not relevant for metals. |
| Photo-oxidative degradation in air | Not relevant for metals. |
| Volatilization | Not relevant for metals |

Monitoring data, if available (Annex VI, para. 44)

| Wolffording data, if available (Affilex VI, para. 44) | |
|---|---------------|
| Soil (indicate location and type of study) | Not available |
| Surface water (indicate location and type of study) | Not available |
| Ground water (indicate location and type of study) | Not available |
| Air (indicate location and type of study) | Not available |
| | |

Chapter 5: Effects on Non-target Species

Toxicity data for aquatic species (most sensitive species of each group) (Annex IIA, Point 8.2, Annex IIIA, Point 10.2)

| Acute toxicity to aquatic organisms | No acute toxicity data are presented as the toxicity was evaluated using a SSD based on chronic toxicity data. |
|---|--|
| Chronic toxicity to aquatic organisms in the FRESHWATER COMPARTMENT | SSD result: HC5-50 = 7.8 µg Cu / l as reasonable worst case Freshwater algae and higher plants: Lowest NOEC used in the SSD = 15.7 µg Cu /L (growth of Pseudokirchneriella subcapitata) Highest NOEC used in the SSD = 510.2 µg Cu /L (growth of Chlorella vulgaris) |
| | Freshwater Invertebrates: Lowest NOEC used in the SSD = 4 µg Cu /L (mortality and reproduction of <i>Ceriodaphnia dubia</i>) Highest NOEC used in the SSD = 181 µg Cu /L (reproduction of <i>Daphnia magna</i>) |
| | Freshwater Fishes: Lowest NOEC used in the SSD = 2.2 µg Cu /L (growth of Oncorhynchus mykiss) Highest NOEC used in the SSD = 188 µg Cu /L (mortality of Perca fluviatilis) |
| Chronic toxicity to aquatic organisms in the SEDIMENT COMPARTMENT | SSD result: HC5-50 = 1741 mg Cu/kg OC, corresponding to 87 mg Cu/kg dry weight for a sediment with 5 % O.C.(TGD default value) |
| | Sediment organisms: |
| | Lowest NOEC used in the SSD = 18.3 mg Cu /kg d.w. (growth and reproduction of <i>Tubifex tubifex</i>) |
| | Highest NOEC used in the SSD = 580.9 mg Cu /kg d.w. (survival of <i>Tubifex tubifex</i>) |
| Chronic toxicity to Sewage microorganisms | The lowest reliable observed NOEC value was noted for the inhibition of respiration = 0.23 mg/l |

Effects on earthworms or other soil non-target organisms

| Acute toxicity to soil organisms (Annex IIIA, point XIII.3.2) | No acute toxicity data are presented as the toxicity was evaluated using a SSD based on chronic toxicity data. |
|---|---|
| Chronic toxicity to soil organisms in the TERRESTRIAL COMPARTMENT | SSD result: HC5-50 = 45.6 mg Cu/kg dry weight was used as reasonable worst case value for Europe in absence of site-specific information on soil properties. |
| | Terrestrial higher plants: Lowest NOEC used in the SSD = 18 mg Cu /kg d.w. (Hordeum vulgare) Highest NOEC used in the SSD = 698 mg Cu /kg d.w. (Lycopersicon esculentum) Terrestrial Invertebrates: Lowest NOEC used in the SSD = 8.4 mg Cu /kg d.w. (cocoon |

| Copper oxide |
|--------------|
|--------------|

| production of <i>Eisenia andrei</i>) Highest NOEC used in the SSD = 1460 mg Cu /kg d.w. (reproduction of <i>Falsomia candida</i>) |
|--|
| Soil micro-organisms: Lowest NOEC used in the SSD = 30 mg Cu /kg d.w. (glucose respiration) Highest NOEC used in the SSD = 2402 mg Cu /kg d.w. (maize respiration) |

Effects on terrestrial vertebrates

| Acute toxicity to mammals (Annex IIIA, point XIII.3.3) | No data |
|---|---------|
| Acute toxicity to birds (Annex IIIA, point XIII.1.1) | No data |
| Dietary toxicity to birds (Annex IIIA, point XIII.1.2) | No data |
| Reproductive toxicity to birds (Annex IIIA, point XIII.1.3) | No data |

Effects on honeybees (Annex IIIA, Point XIII.3.1)

| Acute oral toxicity | No data |
|------------------------|---------|
| Acute contact toxicity | No data |

Effects on other beneficial arthropods (Annex IIIA, Point XIII.3.1)

| Laboratory studies | No data |
|--------------------|---------|
| Semi-field studies | No data |
| Field studies | No data |

Bioconcentration (Annex IIA, Point 7.5)

| Bioconcentration factor (BCF) | For the naturally occurring substances such as essential metals as copper, bioaccumulation is complex, and many processes are available to modulate both accumulation and potential toxic impact. Biota regulates their internal concentrations of essential metals through homeostatic control mechanisms (i.e. active regulation, storage). As a result of these processes, at low metal concentrations, organisms accumulate essential metals more actively in order to meet their metabolic requirements than when they are being exposed at higher metal concentrations. |
|-------------------------------|---|
| | As a consequence of homeostatic processes, and unlike many organic substances, the BCF/BAF is not independent of exposure concentrations for metals and it is inversely related to exposure concentrations. Thus, the use of ratios Cbiota/Cwater or Cbiota/Csediments as an overall approach for estimating copper bioconcentration factors is thus not appropriate. |

| Common avida | |
|--------------|--|
| | |
| Copper oxide | |

| Depuration time (DT ₅₀) | Not applicable for metals |
|---|---------------------------|
| (DT ₉₀) | |
| Level of metabolites (%) in organisms accounting for > 10 % of residues | Not applicable for metals |

Copper oxide 22/09/2011

Appendix II: List of Intended Uses

| Object and/or situation | Member State or Country | Product name | Organisms controlled | Formulati | on | | Application | | Applied | amount per | reatment | Remarks: |
|---|----------------------------------|------------------|-------------------------|---------------------------|----------------|-------------------------------------|--------------------------|--|-------------------|-----------------------|--------------------|--|
| (a) | | | (c) | Type (d-f) | Conc. of as | method kind (f-h) | number min max (k) | interval between applications (min) | g as/L min max | water L/m² min max | g as/m² min max | (m) |
| Timber treatment to prevent damage by fungi and insects Use Class (UC) 1, 2, 3, 4-1 and 4-2 | All | Not available | Fungi and insects | SL (soluble concentratte) | 9.5% | Vacuum/ pressure impregnation | one | Not applicable | See remarks | See remarks | See remarks | Typical concentration of product in working solution: 3.33% for use class 3 with Cu loadings of 1.9 kg/m³. 6% for use class 4-1 and 4-2 with Cu loadings of 3.42 kg/m³. Concentration of copper as a % in the ready to use treating solution: 3.17 g Cu/l for use class 3; 5.7 g Cu/l for use class 4. The concentrate product may contain above 9.5%w/w copper, as long as the product is suitably diluted to give the stated copper concentrations in the treatment solution |

⁽a) e.g. biting and suckling insects, fungi, molds; (b) e.g. wettable powder (WP), emulsifiable concentrate (EC), granule (GR)

⁽c) GCPF Codes - GIFAP Technical Monograph No 2, 1989 ISBN 3-8263-3152-4); (d) All abbreviations used must be explained

⁽e) g/kg or g/l;(f) Method, e.g. high volume spraying, low volume spraying, spreading, dusting, drench;

⁽g) Kind, e.g. overall, broadcast, aerial spraying, row, bait, crack and crevice equipment used must be indicated;

⁽h) Indicate the minimum and maximum number of application possible under practical conditions of use;

⁽i) Remarks may include: Extent of use/economic importance/restrictions

Appendix III: List of studies

Data protection is claimed by the applicant in accordance with Article 12.1(c) (i) and (ii) of Council Directive 98/8/EC for all study reports marked "Yes" in the "Data Protection Claimed" column of the table below. Data protection is claimed under Article 12.1(c) (i) or (ii). These claims are based on information from the applicant. It is assumed that the relevant studies are not already protected in any other Member State of the European Union under existing national rules relating to biocidal products. It was however not possible to confirm the accuracy of this information.

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|---|------|--|---|--|
| A 3.1.1 | Weast, R.C. | 1975 | Handbook of Chemistry and Physics. 56th Edition. CRC Press | No | Public |
| A 3.1.3 | Desai, L.S | 1992 | Physical and Chemical Characteristics. Toxicon Corporation. Project No. 91-GR-0028. GLP, Unpublished | Yes | WPCTF |
| A 3.16 | Desai, L.S | 1992 | Physical and Chemical Characteristics. Toxicon Corporation. Project No. 91-GR-0028. GLP, Unpublished | Yes | WPCTF |
| A 3.17 | Desai, L.S | 1992 | Physical and Chemical Characteristics. Toxicon Corporation. Project No. 91-GR-0028. GLP, Unpublished | Yes | WPCTF |
| A 3.4.1 | O'Connor, B.J. and Mullee, D.M. | 2004 | Copper Oxide: Determination of General Physico-Chemical Properties. Safepharm Laboratories. Project No. 1645/006. GLP, Unpublished | Yes | WPCTF |
| A 3.4.2 | O'Connor, B.J. and Mullee, D.M. | 2004 | Copper Oxide: Determination of General Physico-Chemical Properties. Safepharm Laboratories. Project No. 1645/006. GLP, Unpublished | Yes | WPCTF |
| A 3.5 | O'Connor, B.J. and Mullee, D.M. | 2004 | Copper Oxide: Determination of General Physico-Chemical Properties. Safepharm Laboratories. Project No. 1645/006. GLP, Unpublished | Yes | WPCTF |
| A 3.7 | Desai, L.S | 1992 | Physical and Chemical Characteristics. Toxicon Corporation. Project No. 91-GR-0028. GLP, Unpublished | Yes | WPCTF |
| A 3.9 | Pirot, F., Panisset, F., Agache, P. & Humbert, P. | 1996 | Simultaneous Absorption of Copper and Zinc through Human Skin in vitro. Skin Pharmacol. 9: 43-52. Not GLP, Published. | No | Public |
| A 4.1 | Anonymous | 2000 | Determination of iron, lead and zinc in copper carbonate. Not GLP, unpublished. | Yes | Adchem (Australia) Pty. Limited |
| A 4.1 | S Kibble, M Reed, K Smith | 2005 | Analytical Method for the Determination of Iron, Lead and Zinc in Copper Oxide and Basic Copper Carbonate. , unpublished. | Yes | Adchem (Australia) Pty. Limited |
| A 4.1 | CIPAC | - | CIPAC method for total copper 44/TC/M/3.2. Volumetric thiosulphate method. CIPAC E, Page 44. Not GLP, published. | No | Public |
| A 4.1 | CIPAC | - | CIPAC method for total copper 44/TC/M/3.1. Electrolytic method (Referee method). CIPAC E, Page 42. Not GLP, published. | No | Public |
| A 4.2 | AOAC | 1993 | AOAC Official Method 990.08,. Metals in Solid Wastes; Inductively Coupled Plasma Atomic Emission Method. AOAC Official Methods of Analysis; Metals and Other Elements, Chapter 9, page 31. Not GLP, published. | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|---|------|---|---|--------|
| A 4.2 | EPA | 1983 | Methods for Chemical Analysis of Water and Wastes. Method 220.2 (Copper. Atomic Absorption, furnace technique). Washington, DC; U.S. Environmental Protection Agency. Not GLP, published. | No | Public |
| A 4.2 | EPA | 1983 | Inductively Coupled Plasma – Atomic Emission Spectrometric Method for Trace Element Analysis of Water and Wastes – Method 200.7. Washington, DC; U.S. Environmental Protection Agency. Not GLP, published. | No | Public |
| A 4.2 | EPA | 1986 | Test Methods for Evaluating Solid Waste, Physical/Chemical Methods (SW-846). Method 7210 (Copper. Atomic Absorption, direct aspiration). Washington, DC; U.S. Environmental Protection Agency. Not GLP, published. | No | Public |
| A 4.2 | EPA | 1986 | Methods for Chemical Analysis of Water and Wastes. Method 220.1 (Copper. Atomic Absorption, direct aspiration). Washington, DC; U.S. Environmental Protection Agency. Not GLP, published. | No | Public |
| A 4.2 | NIOSH | 1987 | Method 8005. NIOSH Manual of Analytical Methods, Fourth Edition, 8/15/94. Not GLP, published. | No | Public |
| A 4.2 | NIOSH | 1987 | Method 8310. NIOSH Manual of Analytical Methods, Fourth Edition, 8/15/94. Not GLP, published. | No | Public |
| A 5 | Cockcroft, R. | 1981 | Wood Destroying Basidomyctest Vol. 1. IRG 81/1121 | No | Public |
| A 5 | Connell M, Cornfield J A and Williams G R | 1993 | A New Timber Preservative. Rec of the Annual Convention of the British Wood Preserving and Damp-proofing Association pp 28-36 | No | Public |
| A 5 | Eaton, R.A. & Hale, M.D.C. | 1993 | Wood: Decay Pests and Protection'. Chapman and Hall | No | Public |
| A 5 | Fox R F , Pasek E A , Patel J | 2000 | Laboratory Termite testing of Copper/Boron / Tebuconazole . International Research Group on Wood Preservation. Document No. IRG/WP 00-20192 | No | Public |
| A 5 | Greaves H | 1977 | Potential toxicants for controlling soft rot in hardwoods 1. Laboratory screening tests using a filter paper technique Material und Organismen 12 Bd 1997 Heft | No | Public |
| A 5 | Pohleven, F., Miha, H., Sam, A & Jaka, B. | 2002 | Tolerance of wood decay fungi to commercial copper based wood preservatives. IRG Document No. 02-30291. | No | Public |
| A 5 | Preston A, Walcheski P, Archer K, Zahora A and Jin L | 2000 | The Ground Proximity Decay Test Method, International Research Group on Wood Preservation Doc No. 00-20205 | No | Public |
| A 5 | Price E.A.S and Watson, R.W. | 1962 | Review of water-borne preservatives Rec. of 12th Annual Convention of the British Wood Preserving and Damp-proofing Association, London | No | Public |
| A 5 | Thornton J D | 1977 | Potential toxicants for controlling soft rot in hardwoods II Laboratory tests using sawdust Material und Organismen 12 Bd 1997 Heft 3 | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|--|------|---|---|--------|
| A 6.1.1 | Glaza, S.M. | 2001 | Acute Oral Toxicity Study of Copper Carbonate Dry Light in Rats. Covance Laboratories, Inc. Report No. 7180-100. GLP, Unpublished | Yes | WPCTF |
| A 6.1.1 | Lheritier, M. | 1994 | Test to Evaluate the Acute Toxicity Following a Single Oral Administration (LD50) in the Rat. Pharmakon Europe. Report No. 44193. GLP, Unpublished | Yes | WPCTF |
| A 6.1.1 | Lindena, J | 1990 | Acute Oral Toxicity Test of 'Kupferkarbonat Griin Gefallt 54/56% Cu' in Rats. International Bio Research. Report No. 10- 04-0714-90. GLP, Unpublished | Yes | WPCTF |
| A 6.1.1 | Sanders, A | 2002 | Acute Oral Toxicity Study in the Rat - Acute Toxic Class Method. Safepharm Laboratories. Project No. 1645/001. GLP, Unpublished | Yes | WPCTF |
| A 6.1.2 | Sanders, A | 2002 | Acute Dermal Toxicity Study in the Rat. Safepharm Laboratories. Report No. 1645/002. GLP, Unpublished | Yes | WPCTF |
| A 6.1.4 | Sanders, A | 2002 | Acute Skin Irritation Study in the Rabbit. Safepharm Laboratories. Report No. 1645/003. GLP, Unpublished | Yes | WPCTF |
| A 6.1.4 | Sanders, A | 2002 | Acute Eye Irritation in the Rabbit. Safepharm Laboratories. Report No. 1645/004. GLP, Unpublished | Yes | WPCTF |
| A 6.1.5 | Sanders, A | 2002 | Skin Sensitisation in the Guinea Pig - Magnusson and Kilgman Maximisation Method. Safepharm Laboratories. Report No. 1645/005 . GLP, Unpublished | Yes | WPCTF |
| A 6.12.2 | Chuttani HK, Gupta PS, Gulati S, Gupta DN. | 1965 | Acute Copper Sulfate Poisoning. Am J Med, 39: 849-854; Not GLP; published | No | Public |
| A 6.12.2 | O'Donohue JW, Reid MA, Varghese A, Portmann B, Williams R | 1993 | Micronodular cirrhosis and acute liver failure due to chronic copper self-intoxication. Eur. J. Gastroenterol. 5:561-562; Not GLP; published | No | Public |
| A 6.12.2 | O'Connor, J.M., Bonham, M.P., Turley, E., McKeown, A., McKelvey-Martin, V.J., Gilmore, W.S. and Strain, J.J. | 2003 | Copper supplementation has no effect on markers of DNA damage and liver function in healthy adults (FOODCUE Project). Ann Nutr Metab, 47: 201-206. Not GLP, Published | No | Public |
| A 6.12.2 | Pimentel JC, Marques F | 1969 | Vineyard sprayer's lung - A new occupational disease. Thorax, 2A 4. 678-688; Not GLP; published | No | Public |
| A 6.12.2 | Pimentel JC, Menezes AP. | 1977 | Liver disease in vineyard sprayers. Gastroenterology 72:275-283; Not GLP; published | No | Public |
| A 6.12.2 | Pimentel JC, Menezes AP. | 1975 | Liver granulomas containing copper in vineyard sprayer's lung - A new Etiology of Hepatic Granulomatosis. Am. Rev. Respir. Dis. 111:189-195; Not GLP; published | No | Public |
| A 6.12.2 | Pratt, W.B., Omdahl, J.L. and Sorenson, R.J., | 1985 | Lack of Effects of Copper Gluconate Supplementation. The American Journal of Clinical Nutrition, 42: 681 – 682. Not GLP, Published | No | Public |
| A 6.12.2 | Rock, E., Mazur, A., O'Connor, J.M., Bonham, M.P., Rayssiguier, Y. & | 2000 | The Effect of Copper Supplementation on Red Blood Cell Oxidizability and Plasma Antioxidants in Middle-Aged Healthy Volunteers. Free Radical Biology and | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|--|------|---|---|--------|
| | Strain, J.J | | Medicine. 28 (3); 324-329. Not GLP, Published | (100/110) | |
| A 6.12.2 | Tanner MS, Portmann B, Mowat AP, Williams R, Pandit AN, Mills CF, Bremner I. | 1979 | Increased hepatic copper concentration in Indian Childhood Cirrhosis. Lancet 1:1203-5; Not GLP; published | No | Public |
| A 6.12.2 | Turley, E., McKeown, A., Bonham, M.P., O'Connor, J.M, Chopra, M., Harvey, L.J., Majsak-Newman, G., Fairweather- Tait, S.J., Bugel, S., Sandstrom, B. Rock, E., Mazur, A., Tayssiguier, Y. & Strain, J.J. | 2000 | Copper supplementation in Humans Does Not Affect the Susceptibility of Low Density Lipoprotein to In Vitro Induced Oxidation (Foodcue Project). Free Radical Biology & Medicine, 29: (11); 1129-1134. Not GLP, Published | No | Public |
| A 6.12.4 | Plamenac P, Santic Z, Nikulin A, Serdarevic H. | 1985 | Cytologic changes of the respiratory tract in vineyard spraying workers. Eur J Respir Dis, 67: 50-55; Not GLP; published | No | Public |
| A 6.12.4 | Scheinberg IH, Sternlieb I. | 1994 | Is non-Indian childhood cirrhosis caused by excess dietary copper? Lancet, 344: 1002-1004; Not GLP; published | No | Public |
| A 6.12.4 | Tanner MS, Kantarjian AH, Bhave SA, Pandit AN. | 1983 | Early introduction of copper-contaminated animal milk feeds as a possible cause of Indian Childhood Cirrhosis. Lancet 2: 992-995; Not GLP; published | No | Public |
| A 6.12.5 | International Programme on Chemical Safety | 1990 | Poisons Information Monograph (PIM G002): Copper and copper salts; Not GLP; Published | No | Public |
| A 6.12.7 | International Programme on Chemical Safety | 1990 | Poisons Information Monograph (PIM G002): Copper and copper salts; Not GLP; Published | No | Public |
| A 6.12.8 | International Programme on Chemical Safety | 1990 | Poisons Information Monograph (PIM G002): Copper and copper salts; Not GLP; Published | No | Public |
| A 6.2 | Allen, M.M., Barber, R.S., Braude, R. and Mitchell, K.G. | 1961 | Further studies on various aspects of the use of high-copper supplements for growing pigs. Brit. J. Nutr., 15: 507 – 522, Not GLP Published | No | Public |
| A 6.2 | Amaravadi, R., Glerum, D.M. and Tzagoloff, A. | 1997 | Isolation of a cDNA encoding the human homolog of COX17, a yeast gene essential for mitochondrial copper recruitment. Hum Genet. 99: 329-333. Not GLP, Published. | No | Public |
| A 6.2 | Aoyagi, S. and Baker, D.H. | 1993 | Bioavailability of Copper in Analytical-Grade and Feed Grade Inorganic Copper Sources when Fed to Provide Copper at Levels Below the Chick's Requirement. Poultry Science. 72: 1075-1083. Not GLP, Published | No | Public |
| A 6.2 | Baker, D.H., Odle, J., Funk, M.A. and Wieland, T.M. | 1991 | Research Note: Bioavailability of Copper in Cupric Oxide, Cuprous Oxide, and in a Copper-Lysine Complex. Poultry Science. 70: 177-179. Not GLP, Published | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|---|------|---|---|--------|
| A 6.2 | Buescher, R.G., Griffin, S.A. and Bell, M.C. | 1961 | Copper Availability to Swine from Cu64 Labelled Inorganic Compounds. Journal of Animal Science, 20: 529-531. Not GLP, Published | No | Public |
| A 6.2 | Bunch, R.J., Speer, V.C., Hays, V.W. and McCall, J.T. | 1963 | Effects of High Levels of Copper and Chlortetracycline on Performance of Pigs. J. Animal Sci. 22: 56-60. Not GLP, Published | No | Public |
| A 6.2 | Bunch, R.J., Speer, V.C., Hays, V.W., Hawbaker, J.H. and Catron, D.V. | 1961 | Effects of copper Sulfate, Copper oxide and Chlortetracycline on Baby Pig Performance. J. Animal Sci. 20: 723-726. Not GLP, Published | No | Public |
| A 6.2 | Campbell, C.H., Brown, R. and Linder, M.C. | 1981 | Circulating Ceruloplasmin is an Important Source of Copper for Normal and Malignant Animal Cells. Biochim. Biophys. Acta. 678: 27-38. Not GLP, Published. | No | Public |
| A 6.2 | Cromwell, G.L., Stahly, T.S. and Monegue, H.J. | 1989 | Effects of Source and Level of Copper on Performance and Liver Copper Stores in Weanling Pigs. J. Animal Sci. 67: 2996- 3002. Not GLP, Published | No | Public |
| A 6.2 | Culotta, V.C., Klomp, J.S., Casareno, R.L.B., Krems, B. And Gitlin, J.D | 1997 | The Copper Chaperone for Superoxide Dismutase. The Journal of Biological chemistry. 272 (38): 23469 – 23472. Not GLP, Published | No | Public |
| A 6.2 | Darwish, H.M., Cheney, J.C., Schmitt, R.C. and Ettinger, M.J. | 1984 | Mobilisation of copper (II) from plasma components and mechanism of hepatic copper transport. Am. J. Physiol., 246 (9):G72-G79. Not GLP, Published. | No | Public |
| A 6.2 | Gunshin, H., Mackenzie, B, Berger, U.V., Gunshin, Y., Romero, M.F., Boron, W.F., Nussberger, S., Gollan, J.L. & Hediger, M.A. | 1997 | Cloning and Characterisation of a Mammalian Proton-Coupled Metal-Ion transporter. Nature. 388:482-488. Not GLP, Published. | No | Public |
| A 6.2 | Kegley, E.B. and Spears, J.W. | 1994 | Bioavailability of feed-grade copper sources (oxide, sulfate, or lysine) in growing cattle. J. Animal Sci. 72: 2728-2734. Not GLP, Published | No | Public |
| A 6.2 | Klomp, L.W.J., Lin, S.J., Yuan, D.S., Klausner, R.D., Culotta, V.C. and Gitlin, J.D. | 1997 | Identification and Functional Expression of HAH1, a Novel Human Gene Involved in Copper Homeostasis. The Journal of Biological Chemistry 272(14): 9221-9226. Not GLP, Published. | No | Public |
| A 6.2 | Lee S.H., Lancey R., Montaser A., Madani N., Linder M.C. | 1993 | Ceruloplasmin and copper transport during the latter part of gestation in the rat. Proc Soc Exp Biol Med 203: 428-39. Not GLP, Published. | No | Public |
| A 6.2 | Linder M.C., Weiss K.C. and Hai, V.M. | 1987 | Structure and function of transcuprein in transport of copper by mammalian blood plasma. In: Hurley L.C., Keen C.L., Lonnerdal, B. and Rucker, R.B. (eds). Trace Elements in Man and Animals (TEMA-6). New York: Plenum, 141–144. Not GLP, Published. | No | Public |
| A 6.2 | McArdle, H.J., Gross S.M., Danks D.M. & Wedd, A.G. | 1990 | Role of Albumin's Copper Binding Site in Copper Uptake by Mouse Hepatocytes. Am. J. Physiol. 258 (Gastrointest. Liver Physiol. 21): 988-991. Not GLP, Published. | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|--|------|---|---|--------|
| A 6.2 | McArdle, H.J., Gross, S.M. and Danks, D.M. | 1988 | Uptake of Copper by Mouse Hepatocytes. Journal of Cellular Physiology, 136: 373-378. Not GLP, Published. | No | Public |
| A 6.2 | Norvell, M.J., Gable, D.A. and Thomas, M.C., | 1975 | Effects of feeding high levels of various copper salts to broiler chickens. In Trace Substances in EnvironmentsI Health – 9, (Hemphill, D.D., Ed). University of Missouri, Columbia, MO. Not GLP, Published | No | Public |
| A 6.2 | Pirot, F., Millet, J., Kalia, Y.N. & Humbert, P | 1996 | In vitro Study of Percutaneous Absorption, Cutaneous Bioavailability and Bioequivalence of Zinc and Copper from Five Topical Formulations. Skin Pharmacol. 9: 259-269. Not GLP, Published. | No | Public |
| A 6.2 | Pirot, F., Panisset, F., Agache, P. & Humbert, P. | 1996 | Simultaneous Absorption of Copper and Zinc through Human Skin in vitro. Skin Pharmacol. 9: 43-52. Not GLP, Published. | No | Public |
| A 6.2 | Rojas, L.X., McDowell, L.R., Cousins, R.J., Martin, F.G., Wilkinson, N.S., Johnson, A.B. and Velasquez, J.B. | 1996 | Interaction of different organic and inorganic zinc and copper sources fed to rats. J. Trace Elements Med. Biol. 10: 139-144. Not GLP, Published | No | Public |
| A 6.2 | Scott, K.C. & Turnlund, J.R., | 1994 | Compartment Model of Copper Metabolism in Adult Men. J. Nutr. Biochem. 5: 342-350. Not GLP, Published. | No | Public |
| A 6.2 | Turnlund, J.R., Keen, C.L. and Smith, R.G. | 1990 | Copper status and urinary and salivary copper in young men at three levels of dietary copper. Am. J. Clin. Nutr. 51: 658-64.Not GIP, Published. | No | Public |
| A 6.2 | Turnlund, J.R., Ketes, W.R., Peiffer, G.L. and Scott, K.C | 1998 | Copper absorption, excretion and retention by young men consuming low dietary copper determined using the stable isotope 65Cu. Am. J. Clin. Nutr., 67: 1219 – 1225. Not GLP, Published | No | Public |
| A 6.2 | Turnlund, J.R., Keyes, W.R., Anderson, H.L. and Acord, L.L. | 1989 | Copper absorption and retention in young men at three levels of dietary copper by use of the stable isotope 65Cu. Am. J. Clin. Nutr. 49:870-878.Not GLP, Published. | No | Public |
| A 6.2 | Turnlund, J.R., Wada, L., King, J.C., Keyes, W.R. and Lorra, L.A | 1988 | Copper Absorption in Young Men Fed Adequate and Low Zinc Diets. Biological and Trace Element Research, 17: 31 – 41. Not GLP, Published | No | Public |
| A 6.2 | van Berge Henegouwen, G.P., Tangedahl, T.N., Hofmann, A.F., Northfield, T.C. La Russo, N.F. and McCall, J.T., | 1977 | Biliary Secretion of Copper in Healthy Man. Quantitation by an intestinal perfusion technique. Gastroenterology, 72:1228-1231. Not Published. | No | Public |
| A 6.2 | Van den Berg, G.J., Van Wouwe, J.P and Beynen, A.C., | 1990 | Ascorbic Acid Supplementation and Copper Status in Rats. Biological Trace Element Research, 23: 165-172. Not GLP, Published | No | Public |
| A 6.2 | Walker, R.W. | 1982 | The Result of a Copper Bracelet Clinical Trial and Subsequent Studies. P 469 – 478. In: J. R. J. Sorenson (ed) Inflammatory Diseases and Copper: The Metabolic and Therapeutic Roles of Copper and Other Essential Metalloelements in Humans; | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|---|------|---|---|--------|
| | | | Humana Press; Clifton N.J., USA. Not GLP, Published. | (100.100) | |
| A 6.2 | Weiss K.C. & Linder M.C. | 1985 | Copper transport in rats involving a new plasma protein. Am. J. Physiol. 249: E77-88. Not GLP, Published. | No | Public |
| A 6.2 | Whitaker, P. & McArdle, H.J. | 1997 | Iron Inhibits Copper Uptake by Rat Hepatocytes by Down-Regulating the Plasma Membrane NADH Oxidase. In. Fisher, P.W., L'Abbe, M.R., Cockell, K.A. et al. (Eds). Trace Elements in Man and Animals. (TEMA9). NRC Research Press, Ottowa, pp 237-239 | No | Public |
| A 6.2 | Wirth, W.L. and Linder, M.M. | 1985 | Distribution of Copper Among Components of Human Serum. JNCI. 75: 277-284. Not GLP, Published. | No | Public |
| A 6.2 | Xin, Z., Waterman, D.F., Hemken, R.W., Harmon, R.J. and Jackson, J.A | 1991 | Effects of copper sources and dietary cation-anion balance on copper availability and acid-base status in dietary calves. J. Dairy Sci. 74: 3167-3173. Not GLP, Published | No | Public |
| A 6.2 | Zhou, B. & Gitschier, J. | 1997 | hCTR1; A Human Gene for Copper Uptake Identified by Complementation in Yeast. Proc. Natl. Acad. Sci USA. 94:7481-7486. Not GLP, Published. | No | Public |
| A 6.4.1 | Hébert, C.D., | 1993 | NTP Technical Report on toxicity studies of cupric sulphate (CAS No. 7758-99-8) administered in drinking water and feed to F344/N rats and B6C3F1 mice. National Toxicology Program, Toxicity Report Series No. 29, United States Department of Health and Human Services (NIH Publication 93-3352). GLP, Published | No | Public |
| A 6.4.1 | Hébert, C.D., | 1993 | NTP Technical Report on toxicity studies of cupric sulphate (CAS No. 7758-99-8) administered in drinking water and feed to F344/N rats and B6C3F1 mice. National Toxicology Program, Toxicity Report Series No. 29, United States Department of Health and Human Services (NIH Publication 93-3352). GLP, Published | No | Public |
| A 6.5 | Burki, H.R. and Okita, G.T. | 1969 | Effect of oral copper sulfate on 7,12- dimethylbenz(α)anthracene carcinogenesis in mice. Br. J. Cancer Sep; 23(3): 591-596. Not GLP, Published. | No | Public |
| A 6.5 | Carlton, W.W. and Price, P.S. | 1973 | Dietary Copper and the Induction of Neoplasms in the Rat by Acetylaminofluorene and Dimethylnitrosamine. Fd Cosmet. Toxicol. 11: 827-840 (published). | No | Public |
| A 6.5 | De Vries, D.J., Sewell, R.B. and Beart P.M. | 1986 | Effects of Copper on Dopaminergic Function in the Rat Corpus Striatum. Experimental Neurology, 91: 546-558. Not GLP, Published | No | Public |
| A 6.5 | Hall, E.M. and Butt, E.M | 1928 | Experimental Pigment Cirrhosis Due to Copper Poisoning. It's Relation to Hemochromatosis. Archives of Pathology, 6: 1-25. Not GLP, Published | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|--|------|---|---|--------|
| A 6.5 | Hall, E.M. and Mackay, E.M. | 1931 | Experimental Hepatic Pigmentation and Cirrhosis. I. Does Copper Poisoning Produce Pigmentation and Cirrhosis of the Liver? The American Journal of Pathology, 7: 327-342. Not GLP, Published | No | Public |
| A 6.5 | Harrison, J.W.E., Levin, S.E. and Trabin, B. | 1954 | The Safety and Fate of Potassium Sodium Copper Chlorophyllin and Other Copper Compounds. Journal of the American Pharmaceutical Association, 43(12): 722-737. Not GLP, Published. | No | Public |
| A 6.5 | Haywood, S. | 1980 | The Effect of Excess Dietary Copper on the Liver and Kidney of the Male Rat. Journal of Comparative Pathology, 90: 217-232. Not GLP, Published | No | Public |
| A 6.5 | Haywood, S. | 1985 | Copper Toxicosis and Tolerance in the Rat. I - Changes in Copper Content of the Liver and Kidney. Journal of Pathology, 145: 149- 158. Not GLP, Published | No | Public |
| A 6.5 | Haywood, S. and Comerford, B. | 1980 | The Effect of Excess Dietary Copper on Plasma Enzyme Activity and on the Copper Content of the Blood of the Male Rat. Journal of Comparative Pathology, 90: 233-238. Not GLP, Published | No | Public |
| A 6.5 | Haywood, S. and Loughran, M. | 1985 | Copper Toxicosis and Tolerance in the Rat. II. Tolerance – a Liver Protective Adaptation. Liver, 5: 267-275. Not GLP, Published | No | Public |
| A 6.5 | Haywood, S. and Loughran, M. | 1985 | Copper Toxicosis and Tolerance in the Rat. II. Tolerance – a Liver Protective Adaptation. Liver, 5: 267-275. Not GLP, Published | No | Public |
| A 6.5 | Liu, CC.F. and Medeiros, D.M. | 1986 | Excess Diet Copper Increases Systolic Blood Pressure in Rats. Biological Trace Element Research, 9: 15-24. Not GLP, Published | No | Public |
| A 6.5 | Tachibana, K. | 1952 | Pathological Transition and Functional Vicissitude of Liver During Formation of Cirrhosis by Copper. The Nagoya Journal of Medical Science, 15: 108-114. Not GLP, Published | No | Public |
| A 6.5 | Wiederanders, M.D. and Wasdahl, W.W. | 1968 | Acute and Chronic Copper Poisoning in the Rat. The Journal-Lancet, Minneap, 88: 286-291. Not GLP, Published | No | Public |
| A 6.6.1 | Ballantyne, M. | 1994 | Study to Determine the Ability of Copper II Sulphate Pentahydrate to Induce Mutation in Five Histadine-Requiring Strains of Salmonella typhimurim. Hazleton Europe. Report No. 456/31. GLP, Unpublished | Yes | WPCTF |
| A 6.6.4 | Agarwal, K., Sharma, A. and Talukder, G., | 1990 | Clastogenic effects of copper sulphate on the bone marrow chromosomes of mice in vivo. Mutation Research, 243:1-6. Not GLP, Published | No | Public |
| A 6.6.4 | Bhunya, S.P. & Pati, P.C. | 1987 | Genotoxicity of an Inorganic Pesticide, Copper Sulpahte in Mouse in vivo Test System. Cytologia. 52: 801-808. Not GLP, Published. | No | Public |
| A 6.6.4 | Riley, S.E. | 1994 | Copper II Sulphate Pentahydrate: Induction of Micronuclei in the Bone Marrow of Treated Mice. Hazleton Europe. Report No. 456/33.GLP, Unpublished | Yes | WPCTF |
| A 6.6.4 | Tinwell, H. & Ashby, J. | 1990 | Inactivity of Copper Sulphate in a Bone- Marrow Micronucleus Assay. Mutat. Res. 245: 223-226. Not GLP, Published. | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|---|------|--|---|--------|
| A 6.6.4 | Ward, P.J., | 1994 | Copper II Sulphate Pentahydrate: Measurement of Unscheduled DNA Synthesis in Rat Liver Using an in vivo/in vitro Procedure. Hazleton Europe. Report No. 456/32. GLP, Unpublished | Yes | WPCTF |
| A 6.7 | Burki, H.R. and Okita, G.T. | 1969 | Effect of oral copper sulfate on 7,12- dimethylbenz(α)anthracene carcinogenesis in mice. Br. J. Cancer Sep; 23(3): 591-596. Not GLP, Published | No | Public |
| A 6.7 | Carlton, W.W. and Price, P.S., | 1973 | Dietary Copper and the Induction of Neoplasms in the Rat by Acetylaminofluorene and Dimethylnitrosamine. Fd Cosmet. Toxicol. 11: 827-840. Not GLP, Published. | No | Public |
| A 6.7 | Harrison, J.W.E., Levin, S.E. and Trabin, B., | 1954 | The Safety and Fate of Potassium Sodium Copper Chlorophyllin and Other Copper Compounds. Journal of the American Pharmaceutical Association, 43(12): 722-737. Not GLP, Published | No | Public |
| A 6.8.1 | Aulerich, R.J., Ringer, R.K., Bleavins, M.R. and Napolitano, A. | 1982 | Effects of Supplemental Dietary Copper on Growth, Reproductive Performance and Kit Survival of Standard Dark Mink and the Acute Toxicity of Copper to Mink, 55(2): 337 – 343. Not GLP, Published | No | Public |
| A 6.8.1 | Barash, A., Shoham (Schwartz), Z., Borenstein, R. and Nebel, L. | 1990 | Development of Human Embryos in the Presence of a Copper Intrauterine Device. Gynecol. Obstet. Invest., 29: 203-206. Not GLP, Published | No | Public |
| A 6.8.1 | Barlow, S.M., Knight, A.F. and House, I. | 1981 | Intrauterine exposure to copper IUDs and prenatal development in the rat. J. Rep. Fert., 62: 123 – 130 | No | Public |
| A 6.8.1 | Chang, C.C. And Tatum, H.J. | 1973 | Absence of teratogenicity of intrauterine copper wire in rats, hamsters and rabbits. Contraception, 7(5): 413 – 434 | No | Public |
| A 6.8.1 | DiCarlo, F.J | 1980 | Syndromes of Cardiovascular Malformations Induced by Copper Citrate in Hamsters. Teratology 21: 89-101. Not GLP, Published | No | Public |
| A 6.8.1 | Ferm, V.H. and Hanlon, D.P. | 1974 | Toxicity of Copper Salts in Hamster Embryonic Development. Biology of Reproduction, 11: 97-101. Not GLP, Published | No | Public |
| A 6.8.1 | Haddad, D.S., Al- Alousi, L.A. and Kantarjian, A.H. | 1991 | The Effect of Copper Loading on Pregnant Rats and Their Offspring. Functional and Developmental Morphology 1(3): 17-22. Not GLP, Published | No | Public |
| A 6.8.1 | Kasama, T. and Tanaka, H | 1988 | Effects of copper administration on fetal and neonatal mice. J. Nutr. Sci. Vitaminol., 34: 595-605. Not GLP, Published | No | Public |
| A 6.8.1 | Lecyk, M. | 1980 | Toxicity of CuSO4 in mice embryonic development. Zoologica Poloniae, 28(2): 101-105. Not GLP, Published | No | Public |
| A 6.8.2 | Aulerich, R.J., Ringer, R.K., Bleavins, M.R. and Napolitano, A. | 1982 | Effects of Supplemental Dietary Copper on Growth, Reproductive Performance and Kit Survival of Standard Dark Mink and the Acute Toxicity of Copper to Mink, 55(2): 337 – 343. Not GLP, Published | No | Public |
| A 6.8.2 | Chang, C.C. And Tatum, H.J. | 1973 | Absence of teratogenicity of intrauterine copper wire in rats, hamsters and rabbits. Contraception, 7(5): 413 – 434 | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|--|------|---|---|---------------------------------|
| A 6.8.2 | Cromwell, G.L., Monegue, H.J. and Stahly, T.S. | 1993 | Long-term effects of feeding a high copper diet to sows during gestation and lactation. J. Anim. Sci. 71: 2996-3002. Not GLP, Published | No | Public |
| A 6.8.2 | Hébert, C.D., | 1993 | NTP Technical Report on toxicity studies of cupric sulphate (CAS No. 7758-99-8) administered in drinking water and feed to F344/N rats and B6C3F1 mice. National Toxicology Program, Toxicity Report Series No. 29, United States Department of Health and Human Services (NIH Publication 93-3352). GLP, Published | No | Public |
| A 6.8.2 | Hébert, C.D., | 1993 | NTP Technical Report on toxicity studies of cupric sulphate (CAS No. 7758-99-8) administered in drinking water and feed to F344/N rats and B6C3F1 mice. National Toxicology Program, Toxicity Report Series No. 29, United States Department of Health and Human Services (NIH Publication 93-3352). GLP, Published | No | Public |
| A 6.8.2 | Lecyk, M. | 1980 | Toxicity of CuSO4 in mice embryonic development. Zoologica Poloniae, 28(2): 101-105. Not GLP, Published | No | Public |
| A 6.8.2 | Llewellyn, G.C., Floyd, E.A., Hoke, G.D., Weekley, L.B. and Kimbrough, T.D. | 1985 | Influence of dietary aflatoxin, zinc and copper on bone size, organ weight, and body weight in hamsters and rats. Bull. Environ. Contam. Toxicol., 35: 149-156. Not GLP, Published | No | Public |
| A 6.8.2 | Mylchreest E | 2005 | Copper Sulfate Pentahydrate: Multigeneration Reproduction Study in Rats. DuPont Haskell Laboratory for Health and Environmental Sciences. Laboratory Project I.D.: DuPont-14226; GLP; Unpublished | Yes | European Copper Institute |
| A 6.9 | Murthy, R.C., Lal, S., Saxena, D.K., Shukla, G.S., Mohd Ali, M and Chandra, S.V. | - | Effect of Manganese and Copper Interaction on Behaviour and Biogenic Amines in Rats Fed a 10% Casein Diet. Chem. Biol. Interactions, 37: 299 – 308. Not GLP, Published | No | Public |
| A 7.2.3.1 | Craig, W.B. | 1994 | Adsorption/Desorption of Copper Sulphate in Soil. Inveresk Research International. Project Number 371978. GLP, Unpublished | Yes | WPCTF |
| A 7.2.3.1 | Gayles, M.R. | 1994 | A Study of the Absorption/Desorption of Tanalith 3485 Solutions on Soil. Hickson Timber Protection Products, Protection Products Research Department. Report No. W10/179 | Yes | WPCTF |
| A 7.4.1.1 | Buhl, P.C., & Steven, J.H. | 1990 | Comparative toxicity of inorganic contaminants released by placer mining to early life stages of salmonids. Ecotoxicol. Environ. Saf. Vol. 20, 325-342. Not GLP, Published | No | Public |
| A 7.4.1.1 | Howarth, R. S. & Sprague, J. B. | 1978 | Copper lethality to rainbow trout in waters of various hardness and pH. Water Res. Vol. 12, 455-462. Not GLP, Published | No | Public |
| A 7.4.1.2 | Baird, D. J. et al. | 1991 | A comparative study of genotype sensitivity to acute toxic stress using clones of Daphnia magna. Ecotoxicol Envrion. Saf. Vol. 21, 257-265. Not GLP, Published | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|--|------|--|---|---------------------------------|
| A 7.4.1.2 | Dave, G. | 1984 | Effects of copper on growth, reproduction, survival and haemoglobin in Daphnia magna. Comp. Biochem. Physiol. Vol. 78C (2) 439-443. Not GLP, Published | No | Public |
| A 7.4.1.2 | Le Blanc, G. A. | 1982 | Laboratory investigation into the development of resistance of Daphnia magna to environmental pollutants. Environ. Poll. Vol. A27, 309-322. Not GLP, Published | No | Public |
| A 7.4.1.2 | Oikari, A. et al. | 1992 | Acute toxicities of chemicals to Daphnia magna in humic waters. Sci. Total. Environ. Vol. 117/118, 367-377. Not GLP, Published | No | Public |
| A 7.4.1.3 | De Schamphelaere KAC, Janssen CR | 2005 | A unified bioavailability model for predicting copper toxicity to freshwater green microalgae; University of Gent; Draft report – not yet published; not GLP; Unpublished | No | Public |
| A 7.4.1.3 | Garvey, J.E. et al.* | 1991 | Toxicity of Copper to the Green Alga Chlamydomonas reinhardtii (Chlorophyceae) as Affected by Humic Substances of Terrestrial and Freshwater Origin. Aquatic Toxicology. Vol. 19, 89-96. Not GLP, Published | No | Public |
| A 7.4.1.3 | Ghent University.* | 2002 | Chronic algae testing of copper with Chlamydomonas reinhardtii and Chlorella vulgaris. Unpublished data | No | Public |
| A 7.4.1.3 | Heijerick D., Bossuyt B. and Janssen C. | 2001 | EURO-ECOLE Assessment of the Bioavailability and Potential Ecological Effects of Copper in European Surface Waters - Subproject 4:Evaluation and improvement of the ecological relevance of laboratory generated toxicity data; no report number; not GLP; Unpublished | No | European Copper Institute |
| A 7.4.1.3 | Heijerick* | 2002 | Heijerick DG, Bossuyt BTA, Indeherberg M, Mingazinni M, Janssen CR, 2002. Effects of varying physico-chemistry of European surface waters on the copper toxicity to the green algae Pseudokirchneriella subcapitata. Not GLP, Not Published (submitted). | No | Public |
| A 7.4.1.3 | Nyholm, N. * | 1990 | Expression of results from growth inhibition toxicity tests with algae. Arch. Environ Contam. Toxicol. Vol 19 (4) 518- 522. Not GLP, Published | No | Public |
| A 7.4.1.3 | Schafer, H. et al.* | 1994 | Biotests using unicellular algae and ciliates for predicting long-term effects of toxicant. Ecotoxicol. Enviorn. Safe. Vol. 27, 64-81. Not GLP, Published | No | Public |
| A 7.4.1.3 | Teisseire, H. et al.* | 1998 | Toxic Responses and Catalase Activity of Lemna minor Exposed to Folpet, Copper and their Combination. Ecotoxicol. Environ. Saf. 40, 194-200. Not GLP, Published | No | Public |
| A 7.4.1.4 | Cha, D.K., Allen, H.E. & Song, J.S. | - | Effect of Copper on Nitrifying and Heterotrophic Populations in Activated Sludge. Department of Civil and Environmental Engineering, University of Delaware, USA. Not GLP, Unpublished | Yes | European Copper Institute |
| A 7.4.1.4 | Codina, J.C., Muñoz, M.A., Cazorla, F.M., Perez-Garcia, A., Moriñigo, M.A. & De Vicente, A. | 1998 | The Inhibition of Methanogenic Activity from Anaerobic Domestic Sludges as a Simple Toxicity Bioassay. Water Research. 32 (4) 1338-1342. Not GLP, Published | No | Public |

| | 1 | | | · | |
|------------------|---|------|---|-----------------------------------|--------|
| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Protection Claimed (Yes/No) | Owner |
| A 7.4.1.4 | Madoni, P., Davole, D., Gorbim G. & Vescovi, L. | 1996 | Toxic Effect of Heavy Metals on the Activated Sludge Protozoan Community. Water Research. 30 (1) 135-141. Not GLP, Published | No | Public |
| A 7.4.1.4 | Madoni, P., Davoli, D. & Guglielmi, L. | 1999 | Response of Sour and Aur to Heavy Metal Contamination in Activated Sludge. Water Research. 33 (10): 2459-2464. Not GLP, Published | No | Public |
| A 7.4.2 | Ahsanulla, M. & Williams, A.R. | 1991 | Sublethal Effects and Bioaccumulation of Cadmium, Chromium, Copper and Zinc in the Marine Amphipod, Allorchestes compressa. Mar. Biol. Not GLP, Published. | No | Public |
| A 7.4.2 | Amiard, J.C., Amiard-Triquet, C. & Metayer, C. | 1985 | Experimental Study of Bioaccumulation, Toxicity and Regulation of Some Trace Metals in Various Estuarine and Coastal Organisms. Symp. Biologica. Hung. 29; 313- 323 (published). | No | Public |
| A 7.4.2 | Benoit, D.A. | 1975 | Chronic Effects of Copper on Survival, Growth and Reproduction of the Bluegill (Lepomis macrochirus). Trans. Am. Fish. Soc. 104 (2). 353-358. Not GLP, Published. | No | Public |
| A 7.4.2 | Borgmann, U., Norwood, W.P. & Clarke, C. | 1993 | Accumulation, Regulation and Toxicity of Copper, Zinc, Lead and Mercury in Hyalella azteca. Hydrobiologica. 259: 79-89. Not GLP, Published. | No | Public |
| A 7.4.2 | Brown, B. E | 1977 | Uptake of Copper and Lead by a Metal Tolerant Isopod Asellus meridianus. Freshwater Biol. 7: 235-244. Not GLP, Published. | No | Public |
| A 7.4.2 | Brungs, W. A. Leonard, E.N., McKim, J.M. | 1973 | Acute and Long Term Accumulation of Copper by the Brown Bullhead, Ictalurus nebulosus. J. Fish Res. Board. Can. 30: 583-586. Not GLP, Published. | No | Public |
| A 7.4.2 | Calabrese, A., MacInnes, Nelson, D.A, Greig, R.A. & Yevich, P.P. | 1984 | Effects of Long-Term Exposure to Silver or Copper on Growth, Bioaccumulation and Histopathology in the Blue Mussel, Mytilus edulis. Mar. Environ. Res. 11: 253-274. Not GLP, Published. | No | Public |
| A 7.4.2 | Canterford, G.S., Buchanan, A.S. & Ducker, S.C. | 1978 | Accumulation of Heavy Metals by the Marine Diatom Ditylum brightwelli (West) Grunow. Aust. J. Freshwater Res. 29: 613-22. Not GLP, Published | No | Public |
| A 7.4.2 | Djangmah, J.S. & Grove, D.J. | 1970 | Blood and Hepatopancreas Copper in Crangon vulgaris (Fabricius). Comparative Biochemistry and Physiology (published). | No | Public |
| A 7.4.2 | Engel, D.W. & Brouwer, M. | 1985 | Cadmium and Copper Metallothioneins in the American Lobster, Homarus americanus. Environ. Health. Perspect. 66; 87-92 (published). | No | Public |
| A 7.4.2 | George, S.G., Pirie, B.J.S., Cheyne, A.R., Coombs, T.L. & Grant, P.T. | 1978 | Detoxication of Metals by Marine Bivalves. An Ultrastrutural Study of the Compartmentation of Copper and Zinc in the Oyster Ostrea edulis. Marine Biology, 45; 147-156 (published). | No | Public |
| A 7.4.2 | Graney, R.L, Cherry, D.S. & Carins Jr., J. | 1993 | Heavy Metal Indicator Potential of the Asiatic Clam (Corbicula fluminea) in Artificial Stream Systems. Hydrobiological. 102: 81-88. Not GLP, Published. | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|--|------|---|---|--------|
| A 7.4.2 | Kaland, T. Andersen, T. & Hylland, K. | 1993 | Accumulation and Subcellular Distribution of Metals in the Marine Gastropod Nassarius reticulatus. p37-53. In: R. Dallinger and P.S. Rainbow (eds). Ecotoxicology of Metals in Invertebrates Proceedings of the 1st SETAC European Conference. Lewis Publications; Boca Raton, Fl. USA 461, pp (published). | No | Public |
| A 7.4.2 | Kraak, M. H. S., Lavy, D., Peeters, W. H. M. & Davids, C. | 1992 | Chronic Ecotoxicity of Copper and Cadmium to the Zebra Mussel Dreissena polymorpha. Arch. Envrion. Contam. Toxicol. 23: 363-369. Not GLP, Published. | No | Public |
| A 7.4.2 | McLusky, D.S. & Phillips, C. N. K. | 1975 | Some Effects of Copper on the Polychaete Phyllodoce maculata. Estuarine & Coastal Mar. Sci. 3: 103-108. Not GLP, Published. | No | Public |
| A 7.4.2 | Mersch, J., Morhain, E. & Mouvet, C. | 1993 | Laboratory Accumulation and Depuration of Copper and Cadmium in the Freshwater Mussel Dreissena polymorpha and the Aquatic Moss Rhynchostegium riparioides. Chemosphere. 27 (8): 1475-1485. Not GLP, Published. | No | Public |
| A 7.4.2 | Millanovich, F.P., Spies, R., Guram, M.S. & Sykes, E.E. | 1976 | Uptake of Copper by the Polychaete Cirriformia spirabrancha in the Presence of Dissolved Yellow Organic Matter of Natural Origin. Estuarine and Coastal Mar. Sci. 4: 585-588. Not GLP, Published. | No | Public |
| A 7.4.2 | Pesch, C.E. & Morgan, D | 1978 | Influence of Sediment in Copper Toxicity Tests with the Polychaete Neanthes arenaceodentata. Water Research. 12: 747- 751. Not GLP, Published | No | Public |
| A 7.4.2 | Phillips, D. J. H. | 1976 | The Common Mussel Mytilus edulis as an Indicator of Pollution by Zinc, Cadmium, Lead and Copper. I. Effects of Environmental Variables on Uptake of Metals. Mar. Biol. 38: 59-69. Not GLP, Published. | No | Public |
| A 7.4.2 | Rainbow, P.S. | 1985 | Accumulation of Zn, Cu and Cd by Crabs and Barnacles. Estuarine, Coastal Shelf Science. 21; 669-686 (published). | No | Public |
| A 7.4.2 | Rainbow, P.S. & White, S. L. | 1989 | Comparative Strategies of Heavy Metal Accumulation by Crustaceans: Zinc, Copper and Cadmium in a Decapod and Amphipod and a Barnacle. Hydrobiologia 174; 245-262 (published). | No | Public |
| A 7.4.2 | Rainbow, P.S. & White, S. L. | 1989 | Comparative Strategies of Heavy Metal Accumulation by Crustaceans: Zinc, Copper and Cadmium in a Decapod and Amphipod and a Barnacle. Hydrobiologia 174; 245-262 (published). | No | Public |
| A 7.4.2 | Rainbow, P.S. & White, S. L. | 1989 | Comparative Strategies of Heavy Metal Accumulation by Crustaceans: Zinc, Copper and Cadmium in a Decapod and Amphipod and a Barnacle. Hydrobiologia 174; 245-262 (published). | No | Public |
| A 7.4.2 | Rainbow, P.S., Scott, A.G., Wiggins, E.S. & Jackson, R.W. | 1980 | Effect of Chelating Agents on the Accumulation of Cadmium by the Barnacle Semibalanus balanoides, and Complexation of Soluble Cd, Zn and Cu. Marine Ecology. 2; 143-152 (published). | No | Public |
| A 7.4.2 | Riley, J.P. & Roth, I. | 1971 | The Distribution of Trace Elements in Some Species of Phytoplankton Grown in Culture. J. Mar. Biol. Ass. UK. 51: 63-72. Not GLP, | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|-----------------------|--|------|--|---|--------|
| | | | Published | | |
| A 7.4.2 | Roesijadi, G | 1980 | Influence of Copper on the Clam Protothaca staminea: effects on Gills and Occurrence of Copper Binding Proteins. Biol. Bull. 158: 233-247. Not GLP, Published. | No | Public |
| A 7.4.2 | Shuster, C. N. & B.H. Pringle | 1969 | Trace Metal Accumulation by the American Eastern Oyster, Crassostrea virginica. Proc. Nat. Shellfish. Ass. 59: 91-103. Not GLP, Published. | No | Public |
| A 7.4.2 | Shuster, C.N and Pringle, B.H. | 1969 | Effects of Trace Metals on Estuarine Molluscs. Proceedings of the 1st Mid-Atlantic Industrial Waste Conference. November 13- 15, 197. Not GLP, Published | No | Public |
| A 7.4.2 | Solbe, J.F. de L.G. & Cooper, V.A. | 1976 | Studies on the Toxicity of Copper Sulphate to Stone Loach Neomacheilus Barbatulus (L.) in Hard Water. Wat. Res. 10: 523-527. Not GLP, Published. | No | Public |
| A 7.4.2 | Timmermans, K. R. & Walker, P.A. | 1989 | The Fate of Trace Metals During Metamorphosis of Chrionomids (Diptera, Chironomidae). Environmental Pollution. 62; 73-85 (published). | No | Public |
| A 7.4.2 | White, S.L. & Rainbow, P.S. | 1982 | Regulation and Accumulation of Copper, Zinc and Cadmium by the Shrimp Palaemon elegans. Marine Ecology Progress Series. 8; 95-101 (published). | No | Public |
| A 7.4.2 | Winner , R.W. | 1984 | The Toxicity and Bioaccumulation of Cadmium and Copper as Affected by Humic Acid. Aquatic Toxicology. 5: 267-274. Not GLP, Published. | No | Public |
| A 7.4.2 | Young, J.S., Buschbom, R.L., Gurtisen, J.M. & Joyce, S.P. | 1979 | Effects of Copper on the Sabellid Polychaete, Eudistylia vancouveri: I Concentration Limits for Copper Accumulation. Archives of Environmental Contamination and Toxicology. 8: 97-106. Not GLP. Published | No | Public |
| A 7.4.2 | Zaroogian, G.E. & Johnston, M. | 1983 | Copper Accumulation in the Bay Scallop, Argopecten irradians. Arch. Environ. Contam. Toxicol. 12: 127-133. Not GLP, Published. | No | Public |
| A 7.4.3.1 | Collvin, L.* | 1984 | The effect of copper on growth, food consumption and food conversion of perch Perca fluviatilis L. offered maximal food rations. Aquatic Toxicology (Amsterdam) 6: 105-113. Not GLP, Published | No | Public |
| A 7.4.3.1 | Scudder, B. C., J. L. Carter and H. V. Leland* | 1988 | Effects of copper on development of the fathead minnow, Pimephales promelas Rafinesque. Aquatic Toxicology (Amsterdam) 12: 107-124. Not GLP, Published | No | Public |
| A 7.4.3.1 | Solbe, J. F. d. L. G. and V. A. Cooper* | 1976 | Studies on the toxicity of copper sulphate to stone loach Noemacheilus barbatulus (L.) in hard water. Water Research 10: 523-527. Not GLP, Published | No | Public |
| A 7.4.3.1/ 7.4.3.2 | Horning, W. B. and T. W. Neiheisel* | 1979 | Chronic effect of copper on the bluntnose minnow, Pimephales notatus (Rafinesque). Archives of Environmental Contamination and Toxicology 8: 545-552. Not GLP, Published | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|-----------------------|---|------|---|---|--------|
| A 7.4.3.1/ 7.4.3.2 | Job, K.M., A.M. Askew and R.B. Foster.* | 1995 | Development of a water-effect-ratio for copper, cadmium and lead for the Great Works River in Maine using Ceriodaphnia dubia and Salvelinus fontinalis. Bulletin of Environmental Contamination and Toxicology. 54: 29-35. Not GLP, Published | No | Public |
| A 7.4.3.1/ 7.4.3.2 | McKim, J. M. and D. A. Benoit.* | 1971 | Effects of long-term exposures to copper on survival, growth, and reproduction of brook trout (Salvelinus fontinalis). Journal of the Fisheries Research Board of Canada 28: 655-662. Not GLP, Published | No | Public |
| A 7.4.3.1/ 7.4.3.2 | Mount, D. I. and C. E. Stephan.* | 1969 | Chronic toxicity of copper to the fathead minnow (Pimephales promelas) in soft water. Journal of the Fisheries Research Board of Canada 26: 2449-2457. Not GLP, Published | No | Public |
| A 7.4.3.1/ 7.4.3.2 | Mount, D. I.* | 1968 | Chronic toxicity of copper to fathead minnows (Pimephales promelas Rafinesque). Water Research 2: 215-223. Not GLP, Published | No | Public |
| A 7.4.3.1/ 7.4.3.2 | Mudge, J. E., N. T. E., G. S. Jeane, W. Davis and J. L. Hickam* | 1993 | Effect of varying environmental conditions on the toxicity of copper to salmon. pp. 19-33. In: J. W. Gorsuch, F. J. Dwyer, C. G. Ingersoll and T. W. L. Point (eds.). Environmental Toxicology and Risk Assessment: 2nd Volume, ASTM STP 1216. American Society for Testing Materials, Philadelphia, Pennsylvania. Not GLP, Published | No | Public |
| A 7.4.3.1/ 7.4.3.2 | Sauter, S., K. S. Buxton, K. J. Macek and S. R. Petrocelli* | 1976 | Effects of exposure to heavy metals on selected freshwater fish. Toxicity of copper, cadmium, chromium and lead to eggs and fry of seven fish species. Ecological Reseach Series EPA-600/3-76-105. U.S. Environmental Protection Agency, Environmental Research Laboratory, Duluth, Minnesota. Not GLP, Published | No | Public |
| A 7.4.3.1/ 7.4.3.2 | Spehar, R. L. and J. T. Fiandt* | 1985 | Acute and chronic effects of water quality criteria based metal mixtures on three aquatic species. Project Summary EPA/6000/S3-85/074. U. S. Environmental Protection Agency, Environmental Research Laboratory, Duluth, Minnesota. Not GLP, Published | No | Public |
| A 7.4.3.2 | Belanger, S. E. and D. S. Cherry.* | 1990 | Interacting effects of pH acclimation, pH, and heavy metals on acute and chronic toxicity to Ceriodaphnia dubia (Cladocera). Journal of Crustacean Biology 10(2): 225-235. Not GLP, Published | No | Public |
| A 7.4.3.2 | Brungs, W. A., J. R. Geckler and M. Gast.* | 1976 | Acute and chronic toxicity of copper to the fathead minnow in a surface water of variable quality. Water Research 10: 37-43. Not GLP, Published. | No | Public |
| A 7.4.3.2 | Collvin, L* | 1985 | The effects of copper on maximum respiration rate and growth rate of perch, Perca fluviatilis L. Water Research 18(2): 139-144. Not GLP, Published | No | Public |
| A 7.4.3.2 | Marr, J. C. A., J. Lipton, D. Cacela, J. A. Hansen, H. L. Bergman, J. S. Meyer and C. | 1996 | Relationship between copper exposure duration, tissue copper concentration, and rainbow trout growth. Aquatic Toxicology 36: 17-30. Not GLP, Published | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|--|------|--|---|---------------------------------|
| | Hogstrand.* | | | | |
| A 7.4.3.2 | Pickering, Q., W. Brungs and M. Gast.* | 1977 | Effect of exposure time and copper concentration on reproduction of the fathead minnow (Pimephales promelas). Water Research 11(12): 1079-1083. Not GLP, Published | No | Public |
| A 7.4.3.2 | Seim, W. K., L. R. Curtis, S. W. Glenn and G. A. Chapman.* | 1984 | Growth and survival of developing steelhead trout (Salmo gairdneri) continuously or intermittently exposed to copper. Canadian Journal of Fisheries and Aquatic Sciences 41(3): 433-438. Not GLP, Published | No | Public |
| A 7.4.3.4 | Arthur, J. W. and E. N. Leonard.* | 1970 | Effects of copper on Gammarus pseudolimnaeus, Physa integra, and Campeloma decisum in soft water. Journal of the Fisheries Research Board of Canada 27(7): 1277-1283. Not GLP, Published | No | Public |
| A 7.4.3.4 | Belanger, S. E., J. L. Farris and D. S. Cherry* | 1989 | Effects of diet, water hardness, and population source on acute and chronic copper toxicity to Ceriodaphnia dubia. Archives of Environmental Contamination and Toxicology 18: 601-611. Not GLP, published | No | Public |
| A 7.4.3.4 | Cerda, B and Olive, J.H. | 1993 | Effects of Diet on Seven-Day Ceriodaphnia dubia Toxicity Tests; Ohio J. Sci. 93 (3): 44-47, 1993; no report number; not GLP; Published | No | Public |
| A 7.4.3.4 | Deaver, E. and J. H. Rodgers, Jr.* | 1996 | Measuring bioavailable copper using anodic stripping voltammetry. Environmental Toxicology and Chemistry 15(11): 1925-1930. Not GLP, Unpublished | No | Public |
| A 7.4.3.4 | Hatakeyama, S. and M. Yasuno.* | 1981 | A method for assessing chronic effects of toxic substances on the midge, Paratanytarsus parthenogeneticus-effects of copper. Archives of Environmental Contamination and Toxicology 10: 705-713. Not GLP, Published | No | Public |
| A 7.4.3.4 | Heijerick D., Bossuyt B. and Janssen C. | 2001 | EURO-ECOLE Assessment of the Bioavailability and Potential Ecological Effects of Copper in European Surface Waters - Subproject 4:Evaluation and improvement of the ecological relevance of laboratory generated toxicity data; no report number; not GLP; Unpublished | No | European Copper Institute |
| A 7.4.3.4 | Heijerick DG, Bossuyt BTA, Indeherberg M, Mingazinni M, Janssen CR,* | 2002 | Effects of varying physico-chemistry of European surface waters on the copper toxicity to the green algae Pseudokirchneriella subcapitata. Not GLP, Not Published (submitted). | No | Public |
| A 7.4.3.4 | Jop, K.M., A.M. Askew and R.B. Foster.* | 1995 | Development of a water-effect-ratio for copper, cadmium and lead for the Great Works River in Maine using Ceriodaphnia dubia and Salvelinus fontinalis. Bulletin of Environmental Contamination and Toxicology. 54: 29-35. Not GLP, Published | No | Public |
| A 7.4.3.4 | Maund S.J., E.J. Taylor and D. Pascoe.* | 1992 | Population responses of the freshwater amphipod crustacean Gammarus pulex to copper. Freshwater Biology 28: 29-36. Not GLP, Published | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|--|-------|--|---|--------|
| A 7.4.3.4 | Nebeker, A. V., A. Stinchfield, C. Savonen and G. A. Chapman | 1986 | 1986; Effects of copper, nickel and zinc on three species of Oregon freshwater snails; Environmental Toxicology and Chemistry 5(9): 807-811; not GLP; Published | No | Public |
| A 7.4.3.4 | Nebeker, A. V., C. Savonen, R. J. Baker and J. K. McCrady.* | 1984 | Effects of copper, nickel and zinc on the life cycle of the caddisfly Clistoronia magnifica (Limnephilidae). Environmental Toxicology and Chemistry 3: 645-649. Not GLP, Published | No | Public |
| A 7.4.3.4 | Spehar, R. L. and J. T. Fiandt* | 1985 | Acute and chronic effects of water quality criteria based metal mixtures on three aquatic species. Project Summary EPA/6000/S3-85/074. U. S. Environmental Protection Agency, Environmental Research Laboratory, Duluth, Minnesota. Not GLP, Published | No | Public |
| A 7.4.3.4 | Taylor E.J., S.J. Maund and D. Pascoe.* | 1991 | Evaluation of a chronic toxicity test using growth of the insect Chironomus riparius Meigen. In: Bioindicators and Environmental Management. Eds. D.W. Jeffrey and B. Madden. Academic Press, United Kingdom, pp. 343-352. Not GLP, Published | No | Public |
| A 7.4.3.4 | Van Leeuwen, C. J., J. L. Buchner and H. Van Dijk.* | 1988 | Intermittent flow system for population toxicity studies demonstrated with Daphnia and copper. Bulletin of Environmental Contamination and Toxicology 40(4): 496-502. Not GLP, Published | No | Public |
| A 7.4.3.4 | Winner, R. W.* | 1985 | Bioaccumulation and toxicity of copper as affected by interactions between humic acid and water hardness. Water Research 19(4): 449-455. Not GLP, Published | No | Public |
| A 7.4.3.5 | Belanger, S.E., Farris, J.L., Cherry, D.S., & Cairns, Jr., J. | 1990 | Validation of Corbicula fluminea Growth Reductions Induced by Copper in Artificial Streams and River Systems. Can. J. Fish. Aquat. Sci. 47: 904-914. Not GLP, Published | No | Public |
| A 7.4.3.5 | Borgmann, U., Cover, R. & Loveridge, C. | 1980 | Effects of Metals on the Biomass Production Kinetics of Freshwater Copepods. Can. J. Fish. Aquat. Sci. 37: 567-575. Not GLP, Published. | No | Public |
| A 7.4.3.5 | Clements, W.H., Cherry, D.S. & Cairns Jr., J. | 1988 | Structural Alterations in Aquatic Insect Communities Exposed to Copper in Laboratory Streams. Environ. Toxicol. Chem. 7: 715-722. Not GLP, Published | No | Public |
| A 7.4.3.5 | Clements, W.H., Cherry, D.S. & Cairns, Jr., J. | 1989b | The Influence of Copper Exposure on Predator-Prey Interactions in Aquatic Insect Communities. Freshwater Biology. 21: 483-488. Not GLP, Published. | No | Public |
| A 7.4.3.5 | Clements, W.H., Farris, J.L., Cherry, D.S. & Cairns Jr., J. | 1989a | The Influence of Water Quality on Macroinvertebrate Community Responses to Copper in Outdoor Experimental Streams. Aquatic Toxicol. 14: 249-262. Not GLP, Published | No | Public |
| A 7.4.3.5 | Girling, A.E., Pascoe, S., Janssen, C.R., Peither, A. Wenzel, A., Schafer, H., Neumeier, B., Mitchell, G.C., Taylor, E.J., Maund, S.J., Lay, | 2000 | Development of Methods for Evaluating Toxicity in Freshwater Ecosystems. Ecotoxicol. Environ. Safe. 45, 148-176. Not GLP, Published | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|---|-------|--|---|---------------------------------|
| | J.P., Juttner, I., Crossland, N.O., Stephenson, & Persoone, G. | | | | |
| A 7.4.3.5 | Hart, B.T., Currey, N.A. & Jones, M.J. | 1992 | Biogeochemistry and Effects of Copper, Manganese and Zinc Added to Enclosures in Island Billabong, Magela Creek, Northern Australia. Hydrobiologica. 230: 93-134. Not GLP, Published | No | Public |
| A 7.4.3.5 | Havens, K. E | 1994b | Structural and Functional Responses of Freshwater Plankton Community to Acute Copper Stress. Environmental Pollution. 86 (3); 259-266. Not GLP, Published | No | Public |
| A 7.4.3.5 | Havens, K.E. | 1994a | An Experimental Comparison of the Effects of Two Chemical Stressor on a Freshwater Zooplankton Assemblage. Environmental Pollution. 84: 245-251. Not GLP, Published | No | Public |
| A 7.4.3.5 | Hedtke, S. F. | 1984 | Structure and Function of Copper-Stressed Aquatic Microcosms. Aquatic Toxicology. 5: 227-244. Not GLP, Published | No | Public |
| A 7.4.3.5 | Leland, H.V. & Carter, J.L | 1985 | Effects of Copper on Production of Periphyton, Nitrogen Fixation and Processing of Leaf Litter Sierra Nevada California, Stream. Freshwater Biology. 15: 155-173. Not GLP, Published | No | Public |
| A 7.4.3.5 | Leland, H.V. & Carter, J.L. | 1984 | Effects of copper on Species Composition of Periphyton in a Sierra Nevada, California, Stream. Freshwater Biology. 14: 281-296. Not GLP, Published | No | Public |
| A 7.4.3.5 | Leland, H.V. & Kent, E. | 1981 | Effects of Copper on Microfaunal Species Composition in a Sierra Nevada, California Stream. Verh. Internat. Verein. Liminol. 21: 819-829. Not GLP, Published | No | Public |
| A 7.4.3.5 | Leland, H.V., Fend, S.V., Dudley, T.L. & Carter, J.L. | 1989 | Effects of Copper on Species Composition of Benthic Insects in a Sierra Nevada, California Stream. Freshwater Biology. 21: 163-179. Not GLP, Published | No | Public |
| A 7.4.3.5 | Moore, M.V. & Winner, R.W | 1989 | Relative Sensitivity of Cerodaphnia dubia Laboratory Tests and Pond Communities of Zooplankton and Benthos to Chronic Copper Stress. Aquatic Toxicology. 15: 311-330. Not GLP, Published | No | Public |
| A 7.4.3.5 | Schäfers, C. | 2003 | Community Level Study with Copper in Aquatic Microcosms. Fraunhofer Institute for Molecular Biology and Applied Ecology (IME). Fraunhofer Study Number EECU 01. Not GLP, Unpublished | Yes | European Copper Institute |
| A 7.4.3.5 | Taub, F.B., Kindig, A.C., Meador, J.P. & Swartzman, G.L. | 1991 | Effects of 'Seasonal Succession' and Grazing on Copper Toxicity in Aquatic Microcosms. Verh. Internat. Verein. Limnol. 24: 2205-2214. | No | Public |
| A 7.4.3.5 | Winner, R.W. & Owen, H.A. | 1991 | Seasonal Variability in the Sensitivity of Freshwater Phytoplankton Communities to a Chronic Copper Stress. Aquatic. Toxicol. 19: 73-88. Not GLP, Published. | No | Public |
| A 7.4.3.5 | Winner, R.W., Owen, H.A. & Moore, M.V. | 1990 | Seasonal Variability in the Sensitivity of Freshwater Lentic Communities to a Chronic Copper Stress. Aquatic Toxicology. 17: 75- | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|---|------|--|---|---------------------------------|
| | | | 92. Not GLP, Published | | |
| A 7.4.3.5 | Winner, R.W., Van Dyke, J.S., Caris, N. & Farrel, M.P. | 1975 | Response of the Macroinvertebrate Fauna to a Copper Gradient in an Experimentally Polluted Stream. Verh. Internat. Verein. Liminol. 19: 2121-2127. Not GLP, Published | No | Public |
| A 7.4.3.6 | De Schamphelaere K., Roman, Y.E., Nguyen, L.H. and Janssen, C.R. | 2004 | Bioavailability and ecotoxicity of copper in sediments. Ghent University Report prepared for ECI. Report Ref. PRP ENV-05-59; not GLP; Unpublished. | Yes | European Copper Institute |
| A 7.4.3.6 | Milani D., T.B. Reynoldson, U. Borgmann and J. Kolasa. | 2003 | The relative sensitivity of four benthic invertebrates to metals in spiked sediment exposures and application to contaminated field sediment. Env. Tox and Chem, 22, 4, 845-854; not GLP; Published. | No | Public |
| A 7.4.3.6 | Vecchi M., T.B. Reynoldson, A. Pasteris and D. Bonomi. | 1999 | Toxicity of Copper Spiked Sediments to Tubifex tubifex: Comparison of the 28-day reproductive bioassay with an early life stage bioassay. Env. Tox and Chem, 18, 6, 1173-1179; not GLP; Published | No | Public |
| A 7.5.1.1 | Arshad, M. & Frankenberger, W.T.* | 1991 | Effects of Soil Properties and Trace Elements on Ethylene Production in Soils. Soil Science. 151, 5: 377-386. Not GLP, Published | No | Public |
| A 7.5.1.1 | Arshad, M. & Frankenberger, W.T.* | 1991 | Effects of Soil Properties and Trace Elements on Ethylene Production in Soils. Soil Science. 151, 5: 377-386. Not GLP, Published | No | Public |
| A 7.5.1.1 | Bogomolov, D.M., Chen, S.K., Parmalee, R.W, Subler, S. & Edwards, C.A.* | 1996 | An Ecosystem Approach to Soil Toxicity Testing: A Study of Copper Contamination in Laboratory Soil Microcosms. Applied Soil Ecology. 4: 95-105. Not GLP, Published | No | Public |
| A 7.5.1.1 | Bogomolov, D.M., Chen, S.K., Parmalee, R.W, Subler, S. & Edwards, C.A.* | 1996 | An Ecosystem Approach to Soil Toxicity Testing: A Study of Copper Contamination in Laboratory Soil Microcosms. Applied Soil Ecology. 4: 95-105. Not GLP, Published | No | Public |
| A 7.5.1.1 | Bollag, J-M, Barabasz, W.* | 1979 | Effect of Heavy Metals on the Denitrification Process in Soil. J. Environ. Qual. 8: (2); 196-201. Not GLP, Published | No | Public |
| A 7.5.1.1 | Bollag, J-M, Barabasz, W.* | 1979 | Effect of Heavy Metals on the Denitrification Process in Soil. J. Environ. Qual. 8: (2); 196-201. Not GLP, Published | No | Public |
| A 7.5.1.1 | Chang, F-H. & Broadbent, F.E* | 1981 | Influence of Trace Metals on Carbon Dioxide Evolution from a Yolo Soil. Soil Science. 132: 6; 416-421. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Chang, F-H. & Broadbent, F.E* | 1981 | Influence of Trace Metals on Carbon Dioxide Evolution from a Yolo Soil. Soil Science. 132: 6; 416-421. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Chang, F-H. & Broadbent, F.E* | 1982 | Influence of Trace Metals on Some Soil Nitrogen Transformations. J. Environ. Qual. 11: 1; 1-4. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Chang, F-H. & Broadbent, F.E* | 1982 | Influence of Trace Metals on Some Soil Nitrogen Transformations. J. Environ. Qual. 11: 1; 1-4. Not GLP, Published. | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|--|------|---|---|---------------------------------|
| A 7.5.1.1 | Doelman, P. & Haanstra, L* | 1984 | Short Term and Long Term Effects of Cadmium, Chromium, Copper, Nickel, Lead and Zinc on Soil Microbial Respiration in Relation to Abiotic Soil Factors. Plant and Soil. 79; 317-327. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Doelman, P. & Haanstra, L* | 1984 | Short Term and Long Term Effects of Cadmium, Chromium, Copper, Nickel, Lead and Zinc on Soil Microbial Respiration in Relation to Abiotic Soil Factors. Plant and Soil. 79; 317-327. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Doelman, P. & Haanstra, L* | 1986 | Short and Long Term Effects of Heavy Metals on Urease Activity in Soils. Biol. Fertil. Soils. 2; 213-218. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Doelman, P. & Haanstra, L* | 1986 | Short and Long Term Effects of Heavy Metals on Urease Activity in Soils. Biol. Fertil. Soils. 2; 213-218. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Doelman, P. & Haanstra, L.* | 1989 | Short and Long Term Effects of Heavy Metals on Phosphatase Activity in Soils: An Ecological Dose-Response Model Approach. Biol. Fertil. Soils. 8; 235-241.Not GLP, Published. | No | Public |
| A 7.5.1.1 | Doelman, P. & Haanstra, L.* | 1989 | Short and Long Term Effects of Heavy Metals on Phosphatase Activity in Soils: An Ecological Dose-Response Model Approach. Biol. Fertil. Soils. 8; 235-241.Not GLP, Published. | No | Public |
| A 7.5.1.1 | E. Smolders, Oorts, K | 2004 | Development of a predicitive model of bioavailability and toxicity of copper in soils: microbial toxicity; Laboratory of Soil and Water Management, Katholieke Universiteit Leuven; No report number; not GLP; Unpublished | Yes | European Copper Institute |
| A 7.5.1.1 | Frostegård, A, Tunlid, A. & Bååth, E.* | 1993 | Phospholipid Fatty Acid Composition, Biomass and Activity of Microbial Communities from Two Different Soil Types Experimentally Exposed to Different Heavy Metals. App. Environ. Microbiol. 59 (11): 3605-3617. Not GLP, Published | No | Public |
| A 7.5.1.1 | Frostegård, A, Tunlid, A. & Bååth, E.* | 1993 | Phospholipid Fatty Acid Composition, Biomass and Activity of Microbial Communities from Two Different Soil Types Experimentally Exposed to Different Heavy Metals. App. Environ. Microbiol. 59 (11): 3605-3617. Not GLP, Published | No | Public |
| A 7.5.1.1 | Haanstra, L. & Doelman, P.* | 1984 | Glutamic Acid Decomposition as a Sensitive Measure of Heavy Metal Pollution in Soil. Soil Biol. Biochem. 16; 595-600. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Haanstra, L. & Doelman, P.* | 1984 | Glutamic Acid Decomposition as a Sensitive Measure of Heavy Metal Pollution in Soil. Soil Biol. Biochem. 16; 595-600. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Haanstra, L. & Doelman, P.* | 1991 | An Ecological Dose-Response Model Approach to Short and Long Term Effects of Heavy Metals on Arylsulphatase Activity in Soil. Biol. Fert. Soils. 11; 18-23. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Haanstra, L. & Doelman, P.* | 1991 | An Ecological Dose-Response Model Approach to Short and Long Term Effects of Heavy Metals on Arylsulphatase Activity in Soil. Biol. Fert. Soils. 11; 18-23. Not GLP, | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|---|------|--|---|--------|
| | | | Published. | | |
| A 7.5.1.1 | Hemida, S.K., Omar, S.A. & Abdel-Mallek, A.Y.* | 1995 | Microbial Populations and Enzyme Activity in Soil Treated with Heavy Metals. Water, Air. Soil Poll. 95: 13-22. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Hemida, S.K., Omar, S.A. & Abdel-Mallek, A.Y.* | 1995 | Microbial Populations and Enzyme Activity in Soil Treated with Heavy Metals. Water, Air. Soil Poll. 95: 13-22. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Khan, M. and Scullion, J. | 2002 | Effects of metal (Cd, Cu, Ni, Pb or Zn) enrichment of sewage-sludge on soil micro- organisms and their activities. Applied Soil Ecology, 20, 145-155; not GLP; Published | No | Public |
| A 7.5.1.1 | Maliszewska, W., Dec, S., Wierzbicka, H. & Wozniakowska, A.* | 1985 | The Influence of Various Heavy Metal Compounds on the Development and Activity of Soil Micro-Organisms. Environ. Poll. (Series A). 37; 195-215. Not GLP, Published | No | Public |
| A 7.5.1.1 | Maliszewska, W., Dec, S., Wierzbicka, H. & Wozniakowska, A.* | 1985 | The Influence of Various Heavy Metal Compounds on the Development and Activity of Soil Micro-Organisms. Environ. Poll. (Series A). 37; 195-215. Not GLP, Published | No | Public |
| A 7.5.1.1 | Premi, P.R. & Cornfield, A.H.* | 1969 | Effects of Addition of Copper, Manganese, Zinc and Chromium Compounds on Ammonification and Nitrification During Incubation of Soil. Plant and Soil. 31: (2); 345-352. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Premi, P.R. & Cornfield, A.H.* | 1969 | Effects of Addition of Copper, Manganese, Zinc and Chromium Compounds on Ammonification and Nitrification During Incubation of Soil. Plant and Soil. 31: (2); 345-352. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Quraishi, M.S.I & Cornfield, A.H.* | 1973 | Incubation Study of Nitrogen Mineralisation and Nitrification in Relation to Soil pH and Level of Copper (II) Addition. Environ. Poll. 4; 159-163. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Quraishi, M.S.I & Cornfield, A.H.* | 1973 | Incubation Study of Nitrogen Mineralisation and Nitrification in Relation to Soil pH and Level of Copper (II) Addition. Environ. Poll. 4; 159-163. Not GLP, Published. | No | Public |
| A 7.5.1.1 | Saviozzi, A., Levi- Minzi, R., Cardelli, R. & Riffaldi, R* | 1997 | The Influence of Heavy Metals on Carbon Dioxide Evolution from A Typic Xerochrept Soil. Water, Air Soil Poll. 93: 409-417 (published). | No | Public |
| A 7.5.1.1 | Saviozzi, A., Levi- Minzi, R., Cardelli, R. & Riffaldi, R* | 1997 | The Influence of Heavy Metals on Carbon Dioxide Evolution from A Typic Xerochrept Soil. Water, Air Soil Poll. 93: 409-417 (published). | No | Public |
| A 7.5.1.1 | Skujins, J., Nohrstedt, H-O., Odén, S.* | 1986 | Development of a Sensitive Biological Method for the Determination of a Low-Level Toxic Contamination in Soils. Swedish J. Agric. Res. 16; 113-118.Not GLP, Published | No | Public |
| A 7.5.1.1 | Skujins, J., Nohrstedt, H-O., Odén, S.* | 1986 | Development of a Sensitive Biological Method for the Determination of a Low-Level Toxic Contamination in Soils. Swedish J. Agric. Res. 16; 113-118.Not GLP, Published | No | Public |
| A 7.5.1.1 | Speir, T.W., Kettles, H.A., Percival, H.J. & Parshotam, A* | 1999 | Is Soil Acidification the Cause of Biochemical Response when Soils are Amended with Heavy Metal Salts? Soil Biology and Biochemistry. 31: 1953-1961. | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|---------------------------------|--|-------|--|---|--------|
| | | | Not GLP, Published | | |
| A 7.5.1.1 | Speir, T.W., Kettles, H.A., Percival, H.J. & Parshotam, A* | 1999 | Is Soil Acidification the Cause of Biochemical Response when Soils are Amended with Heavy Metal Salts? Soil Biology and Biochemistry. 31: 1953-1961. Not GLP, Published | No | Public |
| A 7.5.1.2/ 7.5.6 | van Gestel, C.A.M., van Dis, W.A., Dirven-van Breeman, E.M., Sparenburg, P.M. & Baerselman, R.* | 1991 | Influence of Cadmium, Copper and Pentachlorophenol on Growth and Sexual Development of Eisenia andrei (Oligochaeta; Annelida). Biology and Fertility of Soils. 12; 117-121. Not GLP, Published. | No | Public |
| A 7.5.1.2/ 7.5.6 | van Gestel, C.A.M., van Dis, W.A., Dirven-van Breeman, E.M., Sparenburg, P.M. & Baerselman, R.* | 1991 | Influence of Cadmium, Copper and Pentachlorophenol on Growth and Sexual Development of Eisenia andrei (Oligochaeta; Annelida). Biology and Fertility of Soils. 12; 117-121. Not GLP, Published. | No | Public |
| A 7.5.1.2/ 7.5.6/ 7.5.2.1 | Spurgeon, D.J. & Hopkin, S.P.* | 1995 | Extrapolation of the Laboratory-Based OECD Earthworm Toxicity Test to Metal-Contaminated Field Sites. Ecotoxicity. 4; 190-205 | No | Public |
| A 7.5.1.2/ 7.5.6/ 7.5.2.1 | Spurgeon, D.J. & Hopkin, S.P.* | 1995 | Extrapolation of the Laboratory-Based OECD Earthworm Toxicity Test to Metal-Contaminated Field Sites. Ecotoxicity. 4; 190-205 | No | Public |
| A 7.5.1.2/ 7.5.6 | Svendsen, C. & Weeks, J.M* | 1997a | Relevance and Applicability of a Simple Earthworm Biomarker of Copper Exposure. I. Links to Ecological Effects in a Laboratory Study with Eisenia andrei. Ecotoxicology and Environmental Safety. 36; 72-79 | No | Public |
| A 7.5.1.2/ 7.5.6 | Svendsen, C. & Weeks, J.M* | 1997a | Relevance and Applicability of a Simple Earthworm Biomarker of Copper Exposure. I. Links to Ecological Effects in a Laboratory Study with Eisenia andrei. Ecotoxicology and Environmental Safety. 36; 72-79 | No | Public |
| A 7.5.1.2/ 7.5.2.1 | Martin, N.A.,* | 1986 | Toxicity of Pesticides to Allolobophora caliginosa (Oligochaeta: Lumbricidae). New Zealand Journal of Agricultural Research. 29; 699-706. Not GLP, Published. | No | Public |
| A 7.5.1.2/ 7.5.2.1 | Martin, N.A.,* | 1986 | Toxicity of Pesticides to Allolobophora caliginosa (Oligochaeta: Lumbricidae). New Zealand Journal of Agricultural Research. 29; 699-706. Not GLP, Published. | No | Public |
| A 7.5.2.1 | Herbert IN, Svendsen C, Hankard PK and Spurgeon DJ | 2004 | Comparison of instantaneous rate of population increase and critical-effect estimates in Folsomia candida exposed to four toxicants. Ecotox. Environ. Safety 57:175-183; not GLP; Published | No | Public |
| A 7.5.2.1 | Kammenga, J.E., Van Koert, P.H.G., Riksen, J.A.G., Korthals, G.W. & Bakker, J.* | 1996 | A Toxicity Test in Artificial Soil based on the Life History Strategy of the Nematode Plectus acuminatus. Environmental Toxicology and Chemistry. 15; 722-727. Not GLP, Published. | No | Public |
| A 7.5.2.1 | Kammenga, J.E., Van Koert, P.H.G., Riksen, J.A.G., Korthals, G.W. & Bakker, J.* | 1996 | A Toxicity Test in Artificial Soil based on the Life History Strategy of the Nematode Plectus acuminatus. Environmental Toxicology and Chemistry. 15; 722-727. Not GLP, Published. | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|--|------|---|---|---------------------------------|
| A 7.5.2.1 | Lock, K., Janssen, C.R. | 2001 | Test designs to assess the influence of Soil Characteristics on the Toxicity of Copper and Lead to the Oligochaete Enchytraeus albidus; Ecotoxicology, 10, 137-144; not GLP; Published | No | Public |
| A 7.5.2.1 | Ma, W-C.,* | 1988 | Toxicity of Copper to Lumbricid Earthworms in Sandy Agricultural Soils Amended with Cu-Enriched Organic Waste Materials. Ecological Bulletins. 39; 53-56. Not GLP, Published. | No | Public |
| A 7.5.2.1 | Ma, W-C.,* | 1988 | Toxicity of Copper to Lumbricid Earthworms in Sandy Agricultural Soils Amended with Cu-Enriched Organic Waste Materials. Ecological Bulletins. 39; 53-56. Not GLP, Published. | No | Public |
| A 7.5.2.1 | P. Criel, K.A.C. De Schamphelaere and C. R. Janssen | 2005 | Development of a predictive model of bioavailability and toxicity of copper in soils Invertebrate toxicity; Laboratory of Environmental Toxicology and Aquatic Ecology, Ghent University; No Report number; not GLP; Unpublished | Yes | European Copper Institute |
| A 7.5.2.1 | Pedersen, M. B., van Gestel, C.A.M. | 2001 | Toxicity of copper to the collembolan Folsomia fimetaria in relation to the age of soil contamination; Ecotoxicology and Environmental Safety, 49, 54-59; not GLP; Published | No | Public |
| A 7.5.2.1 | Pedersen, M. B., van Gestel, C.A.M., Elmegaard, N. | 2000 | Effects of copper on reproduction of two collembolan species exposed through soil, food and water. Environmental Toxicology and Chemistry, 19, 10, 2579-2588; not GLP; Published | No | Public |
| A 7.5.2.1 | Rundgren S. & Van Gestel, C.A.M. | 1988 | Comparison of Species Sensitivity. In: Handbook of Soil Invertebrate Toxicity Tests. pp 41-55 Ed. H. Lokke and C.A.M. Van Gestel. J. Wiley and Sons Ltd, Chichester, UK; not GLP; Published. | No | Public |
| A 7.5.2.1 | Sandifer, R.D & Hopkin, S.P. | 1997 | Effects of Temperature on the Relative Toxicities of Cd, Cu, Pb and Zn to Folsomia candida (Collembola). Ecotoxicology and Environmental Safety. 37; 125-130; not GLP; Published | No | Public |
| A 7.5.2.1 | Spurgeon DJ, Svendsen C, Kille P, Morgan AJ, Weeks JM. | 2004 | Responses of earthworms (Lumbricus rubellus) to copper and cadmium as determined by measurement of juvenile traits in a specifically designed test system. Ecotoxicol Environ Saf. 57 (1), 54-64; not GLP; Published | No | Public |
| A 7.5.2.1 | van Dis, W.A., Van Gestel, C.A.M., and Sparenburg, P.M. | 1988 | Ontwikkeling van een toets ter bepaling van sublethale effecten van chemische stoffen op regenwormen; RIVM expert report 718480002; not GLP; Published | No | Public |
| A 7.5.2.1 | Van Gestel, C.A.M. & Doornekamp, A. | 1998 | Tests of the Oribatid mite Playnothrus peltifer. In Handbook of Soil Invertebrate Toxicity Tests. Ed. H. Lokke and C.A.M. Van Gestel. J. Wiley and Sons Ltd, Chichester, UK. Not GLP, Published. | No | Public |
| A 7.5.2.1 | Van Gestel, C.A.M. & Doornekamp, A. | 1998 | Tests of the Oribatid mite Playnothrus peltifer. In Handbook of Soil Invertebrate Toxicity Tests. Ed. H. Lokke and C.A.M. Van Gestel. J. Wiley and Sons Ltd, Chichester, UK. Not GLP, Published. | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|--|------|--|---|--------|
| A 7.5.2.1 | van Gestel, C.A.M., Van Dis, W.A., Breemen, E.M. & Sparenburg, P.M. | 1989 | Development of a Standardised Reproduction Toxicity Test with the Earthworm Species Eisenia fetida andrei Using Copper, Pentachlorophenol and 2, 4- Dichloroaniline. Ecotoxicol. Environ. Safety. 18: 305-312; not GLP; Published. | No | Public |
| A 7.5.2.1 | Van Gestel, C.A.M., Van Dis, W.A., Breemen, E.M. & Sparenburg, P.M.* | 1989 | Development of a Standardised Reproduction Toxicity Test with the Earthworm Species Eisenia fetida andrei Using Copper, Pentachlorophenol and 2, 4- Dichloroaniline. Ecotoxicol. Environ. Safety. 18: 305-312. Not GLP, Published. | No | Public |
| A 7.5.2.1 | Van Gestel, C.A.M., Van Dis, W.A., Breemen, E.M. & Sparenburg, P.M.* | 1989 | Development of a Standardised Reproduction Toxicity Test with the Earthworm Species Eisenia fetida andrei Using Copper, Pentachlorophenol and 2, 4- Dichloroaniline. Ecotoxicol. Environ. Safety. 18: 305-312. Not GLP, Published. | No | Public |
| A 7.5.2.2 | Ali, N. A., Ater, M., Sunahara, G., Robidoux, P.Y. | 2004 | Phytotoxicity and bioaccumulation of copper and chromium using barley (Hordeum vulgare L.) in spiked artificial and natural forest soils. Ecotoxicology and Environmental Safety Volume 57, Issue 3, March 2004, Pages 363-374; not GLP; Published | No | Public |
| A 7.5.2.2 | Alva, A.K., Graham, J.H. & Tucker, D.P.H.* | 1993 | Role of Calcium in Amelioration of Copper Phytotoxicity for Citrus. Soil Science. 155; 3; 211-218. Not GLP, Published. | No | Public |
| A 7.5.2.2 | Belanger, A., Levesque, M.P. & Mathur, S.P.* | 1987 | The Influence of Variation in Soil Copper on the Yeild and Nutrition of Radishes Grown in Microplots on Two Peat Soils. International Peat Journal. 2; 65-73.Not GLP, Published. | No | Public |
| A 7.5.2.2 | Belanger, A., Levesque, M.P., Mathur, S.P. | 1987 | The influence of variation in soil copper on the yield and nutrition of radishes grown in microplots on two peat soils. International Peat Journal, 2, 65-73; not GLP; Published | No | Public |
| A 7.5.2.2 | Brun LA, Le Corff J, Maillet J. | 2003 | Effects of elevated soil copper on phenology, growth and reproduction of five ruderal plant species. Environ Pollut. 2003;122(3):361-8; not GLP; Published | No | Public |
| A 7.5.2.2 | Chhibba, I. M. Nayyar, V.K. & Takkar, P.N.* | 1994 | Upper Critical Level of Copper in Wheat (Triticum aestivum) Raised on Typic Ustipsamment Soil. Indian Journal of Agricultural Sciences. 64 (5); 285-289 Not GLP, Published. | No | Public |
| A 7.5.2.2 | de Haan, S., Rethfeld, H. & van Driel, W. | 1985 | Acceptable Levels of Heavy Metals (Cd, Cr, Cu, Ni, Pb, Zn) in Soils, Depending on their Clay and Human Content and Cation-Exchange Capacity. Instituut Voor Bodemvruchtbaarheid Haren-Gr. Report No. 0434-6793 (published). | No | Public |
| A 7.5.2.2 | de Haan, S., Rethfeld, H. & van Driel, W.* | 1985 | Acceptable Levels of Heavy Metals (Cd, Cr, Cu, Ni, Pb, Zn) in Soils, Depending on their Clay and Human Content and Cation-Exchange Capacity. Instituut Voor Bodemvruchtbaarheid Haren-Gr. Report No. 0434-6793. Not GLP, Published | No | Public |
| A 7.5.2.2 | Jarvis, S.C. | 1978 | Copper Uptake and Accumulation by Perennial Ryegrass Grown in Soil and Solution Culture. J. Sci. Fd. Agric. 29:12-18 (published). | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|---|-------|---|---|---------------------------------|
| A 7.5.2.2 | Jarvis, S.C.* | 1978 | Copper Uptake and Accumulation by Perennial Ryegrass Growth in Soil and Solution Culture. J. Sci. Food. Agric. 29: 12-18. Not GLP, Published. | No | Public |
| A 7.5.2.2 | Kalyanaraman, S.B. & Sivagurunathan, P.* | 1993 | Effect of Cadmium, Copper and Zinc on the Growth of Blackgram. Journal of Plant Nutrition. 16 (10), 2029-2042. Not GLP, Published. | No | Public |
| A 7.5.2.2 | Kjær, C. & Elmegaard. N. | 1996 | Effects of Copper Sulphate on Black Bindweed (Polygonum convolvulus L.) Ectoxicology and Environmental Safety. 33; 110-117 (published). | No | Public |
| A 7.5.2.2 | Kjær, C. & Elmegaard. N.* | 1996 | Effects of Copper Sulphate on Black Bindweed (Polygonum convolvulus L.) Ectoxicology and Environmental Safety. 33; 110-117. Not GLP, Published | No | Public |
| A 7.5.2.2 | Korthals, G.W., Alexiev, A.D., Lexmond, T.M., Kammenga, J.E. & Bongers, T.* | 1996a | Long Term Effects of Copper and pH on the Nematode Community in an Agrosystem. Environ. Toxicol. Chem. 15 (6): 979-985. Not GLP, Published. | No | Public |
| A 7.5.2.2 | Lexmond, T. M.* | 1980 | The Effect of Soil pH on Copper Toxicity to Forage Maize Grown Under Field Conditions. Neth. J. Agric. Sci. 28: 164-183. Not GLP, Published. | No | Public |
| A 7.5.2.2 | Lexmond, T.M. | 1980 | The effect of soil pH on copper toxicity to forage maize grown under field conditions. Neth. J. Agric. Science 28, 164-183; not GLP; Published | No | Public |
| A 7.5.2.2 | McBride, M.B.* | 2001 | Cupric Ion Activity in Peat Soil as a Toxicity Indicator for Maize. Journal of Environmental Quality. 30; 78-84. Not GLP, Published | No | Public |
| A 7.5.2.2 | Pedersen, M. B., Kjær, C. Elmegaard, N. | 2000 | Toxicity and Bioaccumulation of Copper to Black Bindweed (Fallopia convolvulus) in Relation to Bioavailability and the Age of Soil Contamination. Archives of Environmental Contamination and Toxicology. 39; 431-439 (published). | No | Public |
| A 7.5.2.2 | Pedersen, M. B., Kjær, C. Elmegaard, N.* | 2000 | Toxicity and Bioaccumulation of Copper to Black Bindweed (Fallopia convolvulus) in Relation to Bioavailability and the Age of Soil Contamination. Archives of Environmental Contamination and Toxicology. 39; 431-439. Not GLP, Published. | No | Public |
| A 7.5.2.2 | Rhoads, F.M., Barnett, R.D. & Olson, S.M.* | 1992 | Copper Toxicity and Phosphorus Concentration in 'Florida 502' Oats. Soil Crop Science Society Florida Proceedings. 51; 18-20. Not GLP, Published. | No | Public |
| A 7.5.2.2 | Rhoads, F.M., Barnett, R.D., Olson, S.M | 1992 | Copper toxicity and phosphorous concentration in "Florida 502" oats. Soil and Crop Science of Florida, 51:18-20; not GLP; Published | No | Public |
| A 7.5.2.2 | Rooney, C.P, McGrath, S.P, Zhao, F.J., Davis, M.R.H., Zhang, H. | 2004 | Development of a predicitive model of bioavailability and toxicity of copper in soils: biological endpoints: Plant toxicity and effects of shock on microbial communities; Agricultural and Environment Division, Rothamsted Research; No report number; not GLP; Unpublished | Yes | European Copper Institute |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|---------------------|--|-------|---|---|--------|
| A 7.5.2.2 | Roth, J.A., Wallihan, E.F. & Sharpless, R.G* | 1971 | Uptake by Oats and Soybeans of Copper and Nickel Added to a Peat Soil. Soil Science. 112; 5; 338-342. Not GLP, Published | No No | Public |
| A 7.5.2/ 7.5.2.1 | Sandifer, R.D. & Hopkin, S. P.* | 1996 | Effects of pH on the Toxicity of Cadmium, Copper, Lead and Zinc to Folsomia candida Willem, 1902 (Collembola) in a Standard Laboratory Test System. Chemosphere. 33: 12; 2475-2486. Not GLP, Published. | No | Public |
| A 7.5.2/ 7.5.2.1 | Sandifer, R.D. & Hopkin, S. P.* | 1996 | Effects of pH on the Toxicity of Cadmium, Copper, Lead and Zinc to Folsomia candida Willem, 1902 (Collembola) in a Standard Laboratory Test System. Chemosphere. 33: 12; 2475-2486. Not GLP, Published. | No | Public |
| A 7.5.6 | Augustsson, A. K., & Rundgren, S* | 1998 | The Enchytraeid Cognettia sphagnetorum in Risk Assessment: Advantages and Disadvantages. Ambio. 27; 62-69. Not GLP, Published. | No | Public |
| A 7.5.6 | Bogomolov, D.M., Chen, S.K., Parmalee, R.W, Subler, S. & Edwards, C.A.* | 1996 | An Ecosystem Approach to Soil Toxicity Testing: A Study of Copper Contamination in Laboratory Soil Microcosms. Applied Soil Ecology. 4; 95-105. Not GLP, Published. | No | Public |
| A 7.5.6 | Jaggy, A. & Streit, B.* | 1982 | Toxic Effects of Soluble Copper on Octolasium cyaneum sav. (lumbricidae). Rev. Suisse De Zoologie. 89; 4: 881-889. Not GLP, Published. | No | Public |
| A 7.5.6 | Kahil, M.A., Abdel- Lateif, H.M., Bayoumi, B.M, van Straalen, N.M. & van Gestel, C.A.M.* | 1996b | Effects of Metals and Metal Mixtures on Survival and Cocoon Production of the Earthworm Aporrectodea caliginosa. Pedobiology. 40; 548-556 .Not GLP, Published. | No | Public |
| A 7.5.6 | Khalil, M.A., Abdel- Lateif, H.M., Bayoumi, B. M. & van Straalen, N.M.* | 1996a | Analysis of Separate and Combined Effects of Heavy Metals on the Growth of Aporrectodea caliginosa (Oligochaeta; Annelida), Using the Toxic Unit Approach. Applied Soil Ecology. 4; 213-219.Not GLP, Published. | No | Public |
| A 7.5.6 | Korthals, G.W, Van de Ende, A., Van Megen, H., Lexmond, T.M., Kammenga, J.E. & Bongers, T. | 1996b | Short-Term Effects of Cadmium, Copper, Nickel and Zinc on Soil Nematodes from Different Feeding and Life-History Strategy Groups. App. Soil. Ecology. 4, 107-117. GLP, published | No | Public |
| A 7.5.6 | Korthals, G.W., Alexiev, A.D., Lexmond, T.M., Kammenga, J.E. & Bongers, T.* | 1996a | Long Term Effects of Copper and pH on the Nematode Community in an Agrosystem. Environ. Toxicol. Chem. 15 (6): 979-985. Not GLP, Published. | No | Public |
| A 7.5.6 | Korthals, G.W., van de Ende, A., van Megen, H., Lexmond, T.M., Kammenga, J. & Bongers, T. * | 1996 | Short-Term Effects of Cadmium, Copper, Nickel and Zinc on Soil Nematodes from Different Feeding and Life History Strategy Groups. Applied Soil. Ecology. 4; 107-117. Not GLP, Published. | No | Public |
| A 7.5.6 | Ma, W.C. * | 1982 | The Influence of Soil Properties and Worm Related Factors on the Concentration of Heavy Metals in Earthworms. Pedobiologica. 24; 109-119. Not GLP, Published. | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|---------------------|---|-------|--|---|--------|
| A 7.5.6 | Parmalee, R. W., Wentsel, R.S., Phillips, C.T., Simini, M. & Checkai, R.T.* | 1993 | Soil Microcosm for Testing the Effects of Chemical Pollutants on Soil Fauna Communities and Trophic Structure. Environ. Toxicol. Chem. 12; 1477-1486. Not GLP, Published. | No | Public |
| A 7.5.6 | Scott-Fordsmand, J.J., Krogh P.H. & Weeks, J.M.* | 1997 | Sublethal Toxicity of Copper to a Soil- Dwelling Springtail (Folsomia fimetaria) (Collembola: Isotomidae). Environmental Toxicology and Chemistry, 16: 12; 2538- 2542.Not GLP, Published. | No | Public |
| A 7.5.6 | Svendsen, C. & Weeks, J. M.* | 1997b | Relevance and Applicability of a Simple Earthworm Biomarker of Copper Exposure. II Validation and Applicability Under Field Conditions in a Mesocosm Experiment with Lumbricus rubellus. Ecotoxicol. Environ. Saf. 36, 80-88. Not GLP, Published. | No | Public |
| A 7.5.6/ 7.5.2.1 | Bengtsson G., Gunnarsson, T. & Rundgren, S.* | 1986 | Effects of Metal Pollution on the Earthworm Dendrobaena rubida (Sav) in Acified Soils. Water, Air and Soil Pollution. 28; 361-383 | No | Public |
| A 7.5.6/ 7.5.2.1 | Krogh, P.H. & Axelsen, J.A.* | 1998 | Test on the Predatory Mite Hypoaspis aculeifer Preying on the Collembolan Folsomia fimetaria. In: Handbook of Soil Invertebrate Toxicity Tests. pp 239-251. Ed. H. Lokke and C.A.M. Van Gestel. J. Wiley and Sons Ltd, Chichester, UK. Not GLP, Published. | No | Public |
| A 7.5.6/ 7.5.2.1 | Kula, H. & Larink, O.* | 1997 | Development and Standardisation of Test Methods for the Prediction of Sublethal Effects of Chemicals on Earthworms. Soil Biology and Biochemistry. 29: 3/4; 635-639. Not GLP, Published. | No | Public |
| A 7.5.6/ 7.5.2.1 | Kula, H. & Larink, O.* | 1998 | Tests of the Earthworms Eisenia fetida and Aporrectodea caliginosa. In: Handbook of Soil Invertebrate Toxicity Tests. pp 95-112 Ed. H. Lokke and C.A.M. Van Gestel. J. Wiley and Sons Ltd, Chichester, UK. Not GLP, Published. | No | Public |
| A 7.5.6/ 7.5.2.1 | Ma, W-C.* | 1984 | Sublethal Toxic Effects of Copper on Growth, Reproduction and Litter Breakdown Activity in the Earthworm Lumbricus rubellus, with Observations on the Influence of Temperature and Soil pH. Environmental Pollution (Series A). 33; 207-219. Not GLP, Published. | No | Public |
| A 7.5.6/ 7.5.2.1 | Rundgren S. & Van Gestel, C.A.M | 1998 | Comparison of Species Sensitivity. In: Handbook of Soil Invertebrate Toxicity Tests. pp 95-112 Ed. H. Lokke and C.A.M. Van Gestel. J. Wiley and Sons Ltd, Chichester, UK. Not GLP. (published). | No | Public |
| A 7.5.6/ 7.5.2.1 | Rundgren S. & Van Gestel, C.A.M.* | 1998 | Comparison of Species Sensitivity. In: Handbook of Soil Invertebrate Toxicity Tests. pp 41-55. Ed. H. Lokke and C.A.M. Van Gestel. J. Wiley and Sons Ltd, Chichester, UK. Not GLP, Published. | No | Public |
| A 7.5.6/ 7.5.2.1 | Sandifer, R.D & Hopkin, S.P.* | 1997 | Effects of Temperature on the Relative Toxicities of Cd, Cu, Pb and Zn to Folsomia candida (Collembola). Ecotoxicology and Environmental Safety. 37; 125-130. Not GLP, Published. | No | Public |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|---------------------|--|-------|--|---|------------------------------|
| A 7.5.6/ 7.5.2.1 | Scott-Fordsmand, J.J., Krogh, P.H & Weeks, J.M.* | 2000a | Responses of Folsomia fimetaria (Collembola: Isotomidae) to Copper Under Different Soil Copper Contamination Histories in Relation to Risk Assessment. Environmental Toxicology and Chemistry. 19: 5; 1297-1303 Not GLP, Published | No | Public |
| A 7.5.6/ 7.5.2.1 | Scott-Fordsmand, J.J., Weeks, J.M & Hopkins, S.P.* | 2000b | Importance of Contamination History for Understanding Toxicity of Copper to Earthworm Eisenia fetica (Oligochaeta: Annelida), Using Neutral Red Retention Assay. Environmental Toxicology and Chemistry. 19: 7; 1774-1780. Not GLP, Published. | No | Public |
| A 7.5.6/ 7.5.2.1 | Spurgeon, D.J., Hopkin, S.P. & Jones, D.T.* | 1994 | Effects of Cadmium, Copper, Lead and Zinc on Growth, Reproduction and Survival of the Earthworm Eisenia fetida (Savigny): Assessing the Environmental Impact of Point-Source Metal Contamination in Terrestrial Ecosystems. Environmental Pollution. 84; 123-130. Not GLP, Published. | No | Public |
| B 6.1.1 | Allen D.J. | 1994 | Acute Oral Toxicity Test in the Rat. Safepharm Laboratories Limited, Project Number 577/17. (Unpublished) | Yes | WPCTF |
| B 6.1.2 | Allen D.J. | 1994 | Acute Dermal Toxicity (Limit Test) in the Rat. Safepharm Laboratories Limited. Project Number 577/18. (Unpublished) | Yes | WPCTF |
| B 6.1.3 | Blagden S.M. | 1994 | Acute Inhalation Toxicity Study Four-Hour Exposure (Nose Only) in the Rat. Safepharm Laboratories Limited. Project Number 577/19. (Unpublished). | Yes | WPCTF |
| B 6.2 | Allen D.J. | 1994 | Allen, D.J. (1994). Acute Dermal Irritation Test in the Rabbit. Safepharm Laboratories Limited. Project Number 577/20. (Unpublished). | Yes | WPCTF |
| B 6.3 | Allen D.J. | 1994 | Buehler Delayed Contact Hypersensitivity Study in the Guinea Pig. Safepharm Laboratories Limited. Project Number 577/22. (Unpublished). | Yes | WPCTF |
| B 7.1/01 | van der Zee, M.E. | 2001 | Submersion tests (NEN 7345) of Tanalith 3494 treated timber. Foundation for Timber Research SHR. Report number 98.214; not GLP; Not published. | Yes | Arch Timber Protection |
| B 7.1/02 | Caswell S | 2000 | Determination of the Emission Profile from Tanalith E Treated Timber. Laboratory Metho to Simulate Exposure from Tanalith E (3491) Treated Timber in an Above Ground (Hazard Class 3) Environment. Hickson Timber Products Limited, Technical Centre. Report No. W20/001a; not GLP; Not published. | Yes | Arch Timber Protection |
| B 7.1/03 | Hohman W.J. | 1993 | Assessing the losses of copper and boron from Tanalith 3485 treated timber with the shower test. Foundation for Timber Research SHR. Report number 92.040; not GLP; Not published. | Yes | Arch Timber Protection |
| B 7.1/04 | Lindegaard, B. | 2003 | Test Methods to Evaluate the leaching of Wood Preservatives from Preservative Treated Wood. Phase II Laboratory Tests. Status Report after 1 st Test Round. Danish Technological Institute, Timber. Report No. 1175742-02; not GLP; not | Yes | Arch Timber Protection |

| Reference No. | Author(s) | Year | Title.Source (where different from company), Company, Report No.GLP (where relevant) / (Un)Published | Data Protection Claimed (Yes/No) | Owner |
|------------------|--------------------------|------|--|---|---------------------------------|
| | | | published. | | |
| B 7.1/05 | Caswell S | 2000 | Determination of the emission profile from TANALITH® E treated timber: Laboratory Method to Simulate Exposure of TANALITH® E (3491) Treated Timber in a Water Contact (Hazard Class 4) Environment. Hickson Timber Products Limited – Technical Centre (Unpublished) | Yes | European Copper Institute |
| | ACP | 2001 | Review of copper chrome arsenic: use as an industrial wood preservative. | No | Public |
| | OECD | 2002 | OECD Series on Testing and Assessment Number 32 and OECD Series on Pesticides Number 10, Guidance Notes for Analysis and Evaluation of Repeat-Dose Toxicity Studies, ENV/JM/MONO(2000)18. | No | Public |
| | Ralph A. & McArdle H. | 2001 | Copper metabolism and copper requirements in the pregnant mother, her fetus and children. | No | Public |
| | WHO | 1990 | Environmental Health Criteria 70, Principles for the Safety Assessment of Food Additives and Contaminants in Food. International Programme on Chemical Safety. | No | Public |
| | WHO | 1998 | Copper. Environmental Health Criteria 200. International Programme on Chemical Safety. | No | Public |