

**Guidance on information requirements and chemical safety  
assessment**

**Appendix R7-2 Recommendations for nanomaterials  
applicable to Chapter R7c Endpoint specific guidance**

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## LEGAL NOTE

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### **Guidance on information requirements and chemical safety assessment**

Extracts from Appendix R7-2 Recommendations for nanomaterials applicable to Chapter R7c - Endpoint specific guidance

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## NOTE

Please note that the present document is a proposed amendment to specific extracts **only** of *Appendix R7-2 to Chapter R.7c* of the IR&CSA Guidance.

This document was prepared by the ECHA Secretariat for the purpose of this consultation and includes only the parts open for the current consultation, i.e. :

- Section 2.1.1 Aquatic bioaccumulation
- Section 2.1.2 Effects on terrestrial organisms

[http://echa.europa.eu/documents/10162/13632/appendix\\_r7c\\_nanomaterials\\_en.pdf](http://echa.europa.eu/documents/10162/13632/appendix_r7c_nanomaterials_en.pdf) (version 1.0 published in April 2012).

The numbering and headings of the sub-sections that are displayed in the document for consultation correspond to those used in the currently published guidance document; this will enable the comparison of the draft revised sub-sections with the current text if necessary.

After conclusion of the consultation and before final publication the updated sub-sections will be implemented in the full documents.

1 **DOCUMENT HISTORY**

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Version	Changes	Date
Version 1	First edition	April 2012
Version 2	<ul style="list-style-type: none"><li>• Update of section 2.1.1. on aquatic bioaccumulation, to explain the general limitations of the Kow as the basis for a waiver for nanomaterials and provide advice on the applicability of the available OECD guidelines;</li><li>• Update of section 2.1.2 on Effects on terrestrial organisms to provide advice on spiking methods and use of different metrics.</li></ul>	Xxxx 2017

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## 2 RECOMMENDATIONS FOR ECOTOXICOLOGICAL ENDPOINTS for NANOMATERIALS:

### 2.1 Specific advice for endpoints

#### 2.1.1 Aquatic bioaccumulation

In the Parent ECHA Guidance, section R.7.10.2 describes the REACH Annex IX information requirements for aquatic bioaccumulation and the use of alternative information when measured data are not available. However, the prediction techniques described in the parent guidance and the use of surrogate information (e.g. the octanol-water partitioning coefficient  $K_{ow}$ ) applicable for many classes of organic substances, cannot be used to predict bioaccumulation potential of nanoparticles.

Section R.7.10.3.2 of the parent guidance concerns non-testing data e.g. Quantitative structure-activity relationships (QSARs), BCF models based on log  $K_{ow}$  and grouping approaches for assessing aquatic bioaccumulation. The use of non-testing approaches such as (Q)SAR approaches in addressing data gaps for nanomaterials is still very limited. In addition to this, the use of such *in silico* models for nanomaterials has also yet to be established or accepted. With regard to nanoparticles, it is not possible to make bioaccumulation estimations based on log  $K_{ow}$  or solubility, as explained in the Appendix R7-1 to ECHA Guidance R.7.a [1] Sections 2.2.1, 2.2.2 and 2.2.4. Therefore the use of non-testing approaches for nanomaterials in deriving an assessment of hazard for the environment must be thoroughly scientifically justified. Appendix R.6-1: Recommendations for nanomaterials applicable to the Guidance on QSARs and Grouping provides an approach on how to justify the use of hazard data between nanoforms of the same substance [2].

Section R.7.10.3.4 of the parent Guidance describes other indicators for bioaccumulation potential. This includes a screening approach where potential bioaccumulation can be estimated from the value of the n-octanol/water partition coefficient ( $K_{ow}$ ). Furthermore, REACH Annex IX 9.3.2 column 2 states that, for instance, a value for log  $K_{ow} \leq 3$  could be used as a waiving argument to omit the testing of bioaccumulation in aquatic species. This approach is not appropriate for nanoparticles, as prediction techniques based on equilibrium partitioning do not apply to nanomaterials - as explained in Appendix R7-1 to ECHA Guidance R.7.a Sections 2.2.1, 2.2.2 and 2.2.4. As outlined in OECD 40, the  $K_{ow}$  value is not suitable for predicting bioaccumulation and not an appropriate endpoint for the physicochemical characterization of nanomaterials.

Taking into account the above, waiving the information requirement for bioaccumulation in aquatic species based on log  $K_{ow}$ , log  $K_{oc}$  or other screening methods should always be accompanied with robust technical and scientific justification for the applicability of the test method used.

#### In vivo tests for aquatic bioaccumulation

The parent guidance section R.7.10.3.1. describes OECD TG 305 Bioaccumulation in Fish [3] : Aqueous and Dietary Exposure as an appropriate *in vivo* test method to fulfill the information requirement set for bioaccumulation in aquatic species in Annex IX 9.3.2.

OECD TG 305 is partially applicable for nanomaterials. It is applicable when the dietary exposure route is followed; the aqueous exposure route resulting in a *bioconcentration factor* (BCF) is not applicable for most of the nanoparticles. BCF is the ratio of the concentration of a substance in an organism to the concentration in water, once a steady state has been achieved. For nanoparticles, a BCF cannot be calculated as no equilibrium will be reached between the organism and the water phase [4] a stable aqueous concentration cannot be

1 maintained, and therefore no plateau is reached and only uptake can be assessed (OECD N 40,  
2 [5], [6] and [7]).

3  
4 A new OECD Guidance for assessing the apparent accumulation potential for nanomaterials is  
5 under development. This guidance, when available, will provide information on how to test  
6 nanomaterials via the dietary exposure and on how to measure and quantify the accumulation  
7 potential in fish.

8  
9 In the meantime, if testing of bioaccumulation in aquatic media is not technically feasible,  
10 bioaccumulation in sediment and soil may be considered as a part of WoE in the  
11 bioaccumulation assessment of nanomaterials. OECD TG 315 Bioaccumulation in Sediment  
12 dwelling Benthic Oligochaetes [8] and OECD TG 317 Bioaccumulation in Terrestrial  
13 Oligochaetes [9] are in principle applicable for nanomaterials, but expert judgement will be  
14 required for performing the bioaccumulation tests and interpreting the results ( [7], [10]).  
15 The results of applying these TGs (OECD TG 315 and OECD TG 317), taking into account the  
16 current challenge in testing bioaccumulation of nanomaterials in fish, may be used as WoE in  
17 bioaccumulation assessment. Soil and sediment compartments are considered potential sinks  
18 for nanomaterials and therefore to also be relevant when considering nanomaterial fate in the  
19 environment.

20  
21 Whenever tests for bioaccumulation in fish or sediment and soil organisms are performed, in  
22 order to be considered reliable, the recommendations on sample preparation and ecotoxicity  
23 and fate testing given in Appendix R7-1 to chapter R7a, section 2.1.1. (Sample preparation)  
24 and Appendix R7-1 to R7b, section 2.1 (General advice on how to perform nanomaterials  
25 ecotoxicity and fate testing) should be followed. In addition, test concentrations should be  
26 monitored throughout the whole test duration to account for concentration-specific changes in  
27 dispersion and agglomeration/aggregation characteristics ( [11], [7]).

## 28 29 **2.1.2 Effects on terrestrial organisms**

### 30 **2.1.2.1 Non testing data**

31 In the parent guidance R7c, part a) of Section R.7.11.3.1, the possibility of using non-testing  
32 approaches e.g. QSAR, grouping and the equilibrium partitioning method (EPM) to estimate  
33 terrestrial toxicity is explained. Regarding nanomaterials, estimates based on “partitioning” are  
34 limited to distribution of a substance in molecular form. In the case of nanomaterials, the  
35 partitioning method may underestimate exposure in soil and sediment environments and  
36 overestimate the exposure in water. If the particle size is small, distribution via air may also  
37 occur. There are no estimation methods available for particle distribution, so this has to be  
38 dealt with on a case-by-case basis.

39 The use of non-testing approaches such as (Q)SAR approaches in addressing data gaps for  
40 nanomaterials is very limited at this time. In addition, the use of *in silico* models for  
41 nanomaterials has yet to be established or accepted. Therefore the use of non-testing  
42 approaches for nanomaterials in deriving an assessment of hazard for the terrestrial  
43 environment must be thoroughly and scientifically justified. Appendix R.6-1: Recommendations  
44 for nanomaterials applicable to the Guidance on QSARs and Grouping provides an approach on  
45 how to justify the use of hazard data between nanofoms of the same substance.

### 46 **2.1.2.2 Testing data**

47 General recommendations for Ecotoxicological and fate testing (including this endpoint) are  
48 given in Section 2.1 of the Appendix on nanomaterials for Chapter R7b.

49 Regarding testing on effects on terrestrial organisms, the methods described in the parent  
50 guidance Section R.7.11 are in principle applicable for testing nanomaterials.

51 The application technique in e.g. sample preparation and spiking has been shown to have an  
52 effect on the availability of the nanomaterial and its level of ecotoxicity in soil ( [5]). Therefore  
53 it is essential that the sample preparation and spiking method applied is well justified and  
54 reported in detail, and that the recommendations set in the OECD Guidance Manual for the

1 Testing Manufactured Nanomaterials: OECD's Sponsorship Programme; first revision [12]  
2 (OECD, 2009), Guidance Notes on Sample Preparation and Dosimetry for nanomaterials [13]  
3 and OECD 40 [14] are followed.

4  
5 When performing the test, the test material needs to be homogenously dispersed in the soil.  
6 OECD 40 [14] describes different spiking methods; particles can be dispersed as aquatic  
7 dispersion into soil (wet spiking) or directly into test media (dry spiking), or put onto a carrier  
8 e.g. silica sand or spiked food. The optimal spiking method depends on both the test material  
9 and the test method. It will depend on the physicochemical properties of the nanomaterial, the  
10 target concentration, the medium, and the bioassay method selected, and preliminary data  
11 gathered prior to the test.

12  
13 Nano-specific metrics such as particle number and surface area should in principle be used  
14 when performing the test as well as the mass metric, unless the use of only mass metric can  
15 be justified (see Section 2.1.1 of Appendix to Chapter R7.b).

16  
17 In addition to these recommendations, it should be considered that measurements of the  
18 nanomaterial's concentration (using different metrics) should be monitored throughout the test  
19 at all test concentrations to account for concentration-specific changes in dispersion and  
20 agglomeration/aggregation characteristics if possible ( [11], [7]).

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